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Journal

River Research and Applications, 27(1)

ISSN

1535-1459

Authors

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Publication Date

2011

DOI

10.1002/rra.1335

Peer reviewed

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5 Running Head:

6 Functional Flows Model Applications

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19

20 **Abstract**

21 This study applies the functional flows model that integrates hydrogeomorphic processes and
22 ecological functions to assess physical habitat. Functional flows are discharge values that serve
23 ecological uses. The model was adjusted to evaluate gravel-bed riffle functionality for fall-run

1 Chinook salmon with respect to river rehabilitation on the Mokelumne River and flood-induced
2 channel change on the Yuba River. The goal was to test if differences in ecological performance
3 were traceable to differences in hydrogeomorphic conditions. Ecological functions studied were
4 bed occupation (spawning, incubation, and emergence) and bed preparation (river bed reworking
5 periods)- both reliant on shear stress dynamics. Model outputs included number of days that have
6 functional flows, ranges of functional flows that provide favorable sediment transport stages, and
7 the efficiency of a site to produce functional flows. Statistical significance of results was tested
8 using non-parametric tests. Functional flows analyses before and after geomorphic alteration
9 indicate that river rehabilitation on the lower Mokelumne River increased the number of days with
10 functional flows, while the Yuba River May 2005 flood increased the functional ranges of flows
11 for the test sites. Reach-scale analyses indicated similar ecological performance at reference sites
12 in both rivers. A comparison between both rivers showed that despite a greater geomorphic
13 potential of the Mokelumne River sites to have functional flows, Yuba River sites actually
14 experienced better ecological performance for fall-run Chinook salmon freshwater life stages due
15 to greater flow availability. The functional flows model provided an objective tool to assess
16 changes in ecological functionality at hydrogeomorphically dynamic sites.

17

1 **Keywords:** aquatic habitat evaluation, ecological functions, functional flows, Mokelumne River,
2 gravel augmentation, Yuba River, floods, sediment transport stages

3

4 **1 Introduction**

5 Hydrogeomorphic processes in rivers determine the conditions of the physical habitat
6 where organisms perform their ecological functions (Knighton, 1998; Marcot and Heyden, 2001;
7 Moyle and Check, 2004). Hydrologic and geomorphologic processes at the watershed scale, such
8 as climate change and landscape evolution, determine the amount of water and sediment that
9 move through catchments and into streams (Poff et al. 1997, Richards et al. 2002). Streamflow
10 and sediment interact dynamically at the reach scale (i.e. defined as a portion of the river with
11 length $> 10^2$ channel widths) through hydrologic, hydraulic, and sediment transport processes
12 controlling local water depth, velocity, bed form, and substrate composition (Fig. 1) (Lisle et al.
13 2000, Parker et al. 2003). The rate at which hydrogeomorphic processes at the watershed and
14 reach scales happen vary from long term geologic trends to rapid alterations of water and/or
15 sediment supply (Major and Mark, 2006; Gibbins *et al.*, 2007; May, 2007). Watershed scale
16 events that induce rapid hydrogeomorphic alterations include river engineering projects and
17 convulsive natural events such as floods, volcanic eruptions, storms, hurricanes, wildfires, mass
18 wasting, volcanic eruptions and earthquakes (Knighton, 1998; Major and Mark, 2006; Moody
19 and Kinner, 2006) (Fig. 1). Such events are likely to determine local hydrologic, hydraulic, and
20 geomorphic changes, however the conditions of physical habitat units that make them suitable to
21 be used by organisms to perform their ecological functions have the potential to persist
22 (Maddock et al. 2004, Tipton et al. 2004, Ito et al. 2006).

1 Events that cause rapid hydrogeomorphic changes have dramatic impacts on local habitat
2 conditions (May, 2007). Gravel augmentation and natural floods change channel morphology,
3 substrate composition, hydraulics, and floodplain connectivity (Wheaton *et al.*, 2004a; Major and
4 Mark, 2006). These alterations affect ecological functionality of physical habitat for organisms
5 that interact with the water column and the river bed (Tipton *et al.*, 2004; Ito *et al.*, 2006). Before
6 a morphologic alteration, specific flow magnitudes generate certain water depths and velocities
7 causing specific bed mobility stages that may be functional for an organism's life stage. After a
8 morphologic alteration, the same flow magnitudes may generate higher or lower water depths
9 and velocities causing a dissimilar bed mobility stage and associated ecological functionality.
10 Consequently, the functionality of a specific hydrograph can change in river sections where rapid
11 hydrogeomorphic changes occur. Likewise, sites with different morphologies may behave
12 different in terms of their hydraulics and sediment transport stages, causing differences in
13 ecological functionality.

14 Assessments of flow functionality before and after changes in physical characteristics of
15 the habitat and at sites with different morphologies within a reach allow the identification of the
16 effect of hydrogeomorphic processes and morphology on ecological functionality. The functional
17 flows model was used to address fundamental research questions by analyzing differences in
18 habitat functionality due to gravel augmentation, natural floods, and differences in channel
19 morphology.

20 The overall goal of this paper was to apply the functional flows model, presented in the
21 preceding article labeled as Part 1 (Escobar-Arias and Pasternack, In Press), to analyze
22 ecological functionality under two types of rapid hydrogeomorphic change processes and under
23 different morphologies. Consequently, the model was used in this study to advance the basic

1 scientific understanding of stream ecological response to hydrogeomorphic change, not just
2 report specific metrics for two rivers. The model was tuned for fall-run Chinook salmon
3 (*Oncorhynchus tshawytscha*), a key endangered Pacific Northwest salmon species that is an
4 indicator of ecosystem functionality (Merz et al. 2004, Augerot et al. 2005, Merz and Chan
5 2005). The model was applied, in two significantly different rivers supporting fall-run Chinook
6 salmon in the same region, to observe the effects of hydraulic and morphologic differences on
7 ecological functionality: the narrow, sediment starved, low-flow lower Mokelumne River that
8 has undergone river rehabilitation through gravel augmentation projects, and the wide lower
9 Yuba River, with an abundance of hydraulic mining sediment in the floodplain and a diverse
10 flow regime, which has undergone rapid morphologic changes due to floods (Fig. 2).

11 The overall hypothesis of this paper is that differences in hydrogeomorphic conditions
12 induce changes in ecological performance of the physical habitat, with the specific mechanism
13 embedded into the algorithms defining the functional flow model, as outlined in the preceding
14 article (Escobar-Arias and Pasternack, In Press). Research questions 1 to 3 listed below were
15 formulated to test the hypothesis by comparing functional flows before and after morphologic
16 alterations. Such morphologic alterations were caused by rapid hydrogeomorphic changes at
17 habitat units defined as zones with characteristic physical attributes where organisms perform
18 ecological functions (Knighton 1998, Marcot and Heyden 2001, Moyle and Check 2004).
19 Additional research questions 4 and 5 listed below were posited to compare functional flows
20 among sites within the same river reach and between rivers:

- 21 1) What are the ranges of flows that are potentially functional?
- 22 2) What is the number of days that flows are functional?
- 23 3) What is the efficiency of a habitat unit in producing functional flows?

1 4) What is the overall functionality of habitat units within a river reach as measured by the
2 ranges of functional flows, number of days with functional flows, and efficiency to
3 produce functional flows?

4 5) How does functionality of a reach as measured by the ranges of functional flows, number
5 of days with functional flows, and efficiency to produce functional flows compare to
6 other reaches in the river and to other rivers?

7 In order to answer the research questions, the method used involved 1) performing
8 functional flows analyses for theoretical and actual water years to study the effects of gravel
9 augmentation on ecological functionality on the lower Mokelumne River and to study the effects
10 of flood-induced morphologic changes on ecological functionality on the lower Yuba River and
11 2) comparing results of the functional flows analyses before and after morphologic alteration,
12 among different habitat units within the same river reach, and between the rivers to observe the
13 utility of applying the model in habitat units with different hydrologic regimes and
14 morphologies. The analyses presented in this paper are an example of the use of the functional
15 flows model that highlights the utility of the tool for basic and applied science and management.
16 The application presented shows that the functional flows analysis provides a uniform measure
17 that can be used to assess and compare ecological functionality among habitat units on a single
18 river and among rivers.

20 **2 Functional Flows Analysis**

21 The Functional Flows Model (FFM) uses assessments of hydrogeomorphic dynamics as
22 indicated by the temporal pattern of shear stress, which is a key factor determining physical
23 habitat for several ecological functions. A functional flow is defined as a discharge that interacts

1 with river bed morphology through hydraulic processes providing a shear stress value that serves
2 an ecological use. Depending on the specificity of a given ecological function, functional flows
3 may occur over a range of discharges.

4 The next subsections present an overview of the functional flows analysis. To classify
5 ranges of flows that are functional, the initial step was to identify relevant ecologic functions for
6 the target species, in this case, fall-run Chinook salmon. Subsequently, habitat units within the
7 two river systems were localized. Data for the calculations was gathered based upon the method
8 to estimate shear stress selected (i.e. discharge time series, cross-section geometry, water surface
9 slope, and grain size distribution). The final task was to build a table of functionality to specify
10 the dependence of ecological functions on river bed sediment transport stages determined by
11 shear stress thresholds and to assign functionality to streamflow time series.

12 **2.1 Ecological Functions**

13 Ecological functions are defined as the ways in which organisms interact with and use
14 their physical habitat (Marcot and Heyden, 2001). Ecological functions related to physical
15 habitat happening during the freshwater life stage of salmon include upstream migration of
16 adults, spawning, embryo incubation, fry emergence, and juvenile rearing. Every year salmon
17 migrate to upstream reaches to spawn in foothill and mountain cold water streams (Reiser and
18 Bjornn, 1979). They initiate the construction of the nest, called redd, by digging a hole to depths
19 that vary depending on the size classes of the females for each species (DeVries 1997,
20 Montgomery et al. 1999). After females lay their eggs and males fertilize them, the females
21 finish the redd construction by covering the embryos with gravel (Groot and Margolis, 1991;
22 DeVries, 1997). During incubation, the embryos remain buried within the gravel. After a period
23 that ranges from 2-8 months, the just-hatched fish, called fry, emerge through the gravel to begin

1 their juvenile life in freshwater (Groot and Margolis 1991, Merz et al. 2004, Augerot et al. 2005).
2 To simplify the analysis, ecological functions of interest are grouped as 1) bed occupation
3 functions that occur in periods when the fish interact with the river bed (i.e. spawning,
4 incubation, and emergence), and 2) bed preparation functions that modify river bed surface
5 conditions for the next spawning season (Fig. 3A). High and low flows may be functional or not
6 depending on the timing with respect to the selected ecological functions (Fig. 3B).

7 **2.2 Habitat Units**

8 Physical habitat units in rivers are zones with characteristic physical attributes where
9 organisms perform ecological functions (Knighton, 1998; Marcot and Heyden, 2001; Moyle and
10 Check, 2004). The selection of sites for functional flows applications concentrates on sites where
11 the ecological functions under analysis are expected to occur. Preferred spawning habitat units
12 are areas with low water depths, moderate velocities, and gravel that fish can move for redd
13 construction (Lisle and Lewis, 1992; Kondolf and Wolman, 1993; DeVries, 1997; Gallagher and
14 Gard, 1999; Lapointe *et al.*, 2000; Moir *et al.*, 2004). Flow, bed topography, and sediment
15 sorting at the pool tail provide the bed form and water depth and velocity that salmon seek to
16 carry out their reproductive life stage (Emery *et al.*, 2003). Consequently, pool tail/riffle entrance
17 is one preferred location for spawning (Montgomery et al. 1999, Coulombe-Pontbriand and
18 LaPointe 2004, Moir et al. 2004, Moyle and Check 2004). Other locations include side channels
19 and lateral bars (Webb et al. 2001, Moir et al. 2004, Morley et al. 2005).

20 **2.3 Equations for Shear Stress Calculation**

21 Shear stress is the key parameter that represents the force available to scour the river bed
22 and can be used to delimit ecological functions that are highly dependent on sediment transport

1 regimes at selected habitat units (Montgomery *et al.*, 1999). The 1D Saint Venant equation for
 2 non-steady non-uniform flow or at least 9 other methods could be used to estimate boundary
 3 shear stress from field measurements (Dietrich and Whiting, 1989) for functional flows analysis.
 4 For the application presented in this study, the simplified depth-slope product was selected to
 5 calculate boundary shear stress

$$6 \quad \tau_0 = \rho g R S \quad \text{or} \quad \tau_0 = \rho g h S \quad \text{for wide channels} \quad (1)$$

7 This method assumes conditions of uniform and steady flow that need to be checked for
 8 actual applications. Eq. 1 is used as a first cut approach that incorporates the dominant hydraulic
 9 interactions controlling channel sediment transport (Konrad *et al.*, 2002; Buffington *et al.*, 2004;
 10 Murray, 2007). Using this simplified expression for shear stress it was possible to focus this
 11 study on exploring the interactions among physical processes and ecological functions without
 12 the need for calculating shear stress in detail, which is a valuable effort that has been the focus of
 13 several studies (Booker, 2003; Rodriguez *et al.*, 2004; Wilson *et al.*, 2006).

14 “At-a-station” cross-section geometry relations can be used to evaluate depth at a range
 15 of discharge values

$$16 \quad h = c Q^f \quad (2)$$

17 where c and f are empirical values that control the water depth response to discharge increments
 18 at the cross section (Leopold and Maddock, 1953; Parker, 1979). This approach is useful to
 19 obtain water depth time series necessary to estimate shear stress time series.

20 Replacing Eq. 2 in Eq. 1, the shear stress becomes

$$21 \quad \tau_0 = \rho_w g (c Q^f) S \quad (3)$$

22 Eq. 3 can be non-dimensionalized to obtain non-dimensional absolute values of shear
 23 stress

$$\tau_0^* = \frac{\tau_0}{g(\rho_s - \rho)D_{50}} \quad (4)$$

where ρ_s is the sediment density. Functional flows are expressed in terms of non-dimensional shear stress τ_0^* , which allows a generalized definition of the model. Non-dimensional boundary shear stress can be compared to non-dimensional absolute values of τ_0^* that represent the critical magnitude necessary to entrain gravel of a given size, τ_{crit}^* , or Shields parameter (Buffington and Montgomery 1997, Wheaton et al. 2004a). Substituting Eq. 2 and 3 into Eq. 4, a new form of τ_0^* is obtained

$$\tau_0^* = \frac{\rho ghS}{g(\rho_s - \rho)D_{50}} = \frac{\rho(cQ^f)S}{(\rho_s - \rho)D_{50}} \quad (5)$$

Eq. 5 can be used to evaluate τ_0^* for discharge time series and for a given cross section with a specific median grain size. The temporal pattern of shear stress represents geomorphic dynamics that are relevant to identify functional transport stages for fall-run Chinook salmon life stages (Fig. 4).

2.4 Shear Stress Thresholds

The FFM requires specification of bed mobility transport stages delimited by boundary shear stress thresholds for selected ecological functions (Table 1, Column 1) (Kondolf and Wilcock, 1996; Lisle *et al.*, 2000). Bed mobility categories are high flow/full mobility (FM), intermediate high flow/interstitial fines mobility (IFM), intermediate low flow/ superficial fines mobility (SFM), and low flow/stable bed (SB) (Fig. 4). Associated dimensionless critical shear stress values are used to delimit bed mobility stages for gravel-bed rivers according to values found in the literature (Table 1, Column 2B). For the present application of functional flows analysis, a stable bed is assumed when $\tau_0^* < 0.01$, intermittent transport when $0.01 < \tau_0^* < 0.03$,

1 partial transport when $0.03 < \tau_0^* < 0.06$, and full mobility when $0.06 < \tau_0^* < 0.10$ (Buffington and
 2 Montgomery 1997, Lisle et al. 2000). The upper threshold for full mobility is set at 0.10
 3 assuming that beyond this point there is intensive bed load transport that is non-functional to
 4 support spawning functions (Lisle *et al.*, 2000). Further details on the thresholds for bed mobility
 5 stages are provided in the preceding paper labeled as Part 1 (Escobar-Arias and Pasternack, In
 6 Press).

7 **2.5 Model Structure and Table of Functionality**

8 The FFM algorithm integrates key relations between shear stress and ecological functions
 9 that have already been identified and are available in the literature such as the ones presented
 10 above. By estimating τ_0^* as a function of discharge time series it is possible to create the “table
 11 of functionality” to determine the functionality of sediment transport stages and flows serving
 12 ecological functions (Table 1).

13 In addition to temporal changes in bed mobility stages represented by Eq. 5, it is possible
 14 to observe the dependence of the geomorphic dynamics on streamflow. Q can be non-
 15 dimensionalized by a combination of variables with length and time dimensions (i.e. $L^{-3}T^1$). The
 16 formulation by Parker et al. (1979) is used

$$17 \quad Q^* = \frac{Q}{\sqrt{gD_{50}D_{50}^2}} \quad (6)$$

18 Eqs. 5 and 6 can be used to produce curves of τ_0^* vs. Q^* to observe shear stress as a
 19 function of streamflow (Fig. 5). Curves of non-dimensional quantities allow comparison of
 20 channels with a wide range of characteristics and have been used to group and observe trends in
 21 data of rivers from different geographic regions (Parker *et al.*, 2003). In this study, the resultant
 22 curve τ_0^* vs. Q^* , where τ_0^* is function of S , D_{50} , c and f ; and Q^* is function of Q and D_{50} , depicts

1 the variation in bed mobility stages for a cross section with a particular slope, median grain size,
2 and geometry for a range of discharges. Each portion of the curve within thresholds of τ_{c50}^* (0.01,
3 0.03, 0.06, and 0.1) can be categorized as functional or non-functional. The model was
4 programmed in MATLAB to facilitate calculations. The MATLAB code is available for public
5 use by requesting it from the authors.

6 The FFM provides an approach to understand the relations among hydrogeomorphic
7 parameters and ecological functions based on a representation of the natural system (Murray,
8 2007). The analytical algorithm does not provide a predictive model, but an explanatory model
9 conceived to explain links among physical processes and biological systems (van Asselt and
10 Rotmans 2002, Murray 2003). Inputs of the model including discharge (Q), a parameter (f)
11 governing depth response to incremental discharge changes, water surface slope (S), and median
12 grain size (D_{50}) are large-scale parameterizations of the interactions among the most significant
13 variables representing physical and ecological processes. Outputs of the model including
14 functional ranges of Q^* , number of days with functional flows, and percent efficiency for a
15 habitat unit and for a water year permit measuring such interactions under different sets of
16 conditions resulting in data that adds to the existing knowledge about hydrogeomorphic and
17 ecologic links. These results are derived from relations among streamflow time series, water
18 depth, and shear stress that have been tested. The validation of these results would require
19 extensive measures of discharge values, sediment transport stages, and ecological functions at all
20 sites all days within a water year. Therefore, the functional flows model constitutes a theoretical
21 analysis with scientific basis and its actual validation is beyond the objective of the present study
22 (Murray, 2003).

23

3 Field Sites

The Mokelumne and Yuba Rivers flow generally west, draining watersheds covering 1,624 km² and 3,480 km² respectively, of the central Sierra Nevada of California (Figure 2). The Mokelumne River is a tributary to the San Joaquin River, while the Yuba River is a tributary to the Feather River. Their watersheds are ~110 km apart. The upstream reaches of both rivers receive ~1,200 mm of precipitation annually, while the central region of the watersheds receives ~510 mm (Mount, 1995). Water feeds the rivers mostly as rainfall runoff in the late fall and winter, and then as snowmelt in the spring. Historically, both rivers have been manipulated for dam construction, gold mining, gravel extraction, hydropower generation, water supply, and flood regulation (Pasternack *et al.*, 2004; Pasternack, 2008b). Such anthropogenic influence has caused in-stream physical habitat degradation (Mount, 1995). Recently, both rivers have undergone morphological alterations produced by dissimilar causes: artificial gravel augmentation on the lower Mokelumne River (Wheaton *et al.*, 2004b; Elkins *et al.*, 2007), and natural floods on the lower Yuba River (Pasternack, 2008b). These events have modified channel form and habitat conditions.

3.1 Gravel Augmentation on the lower Mokelumne River

The Mokelumne River has 16 major impoundments, the two largest being Pardee Reservoir (259 million m³), completed in 1929, and Camanche Reservoir (531 million m³), completed in 1963 (Pasternack *et al.*, 2004). East Bay Municipality Utility District (EBMUD) manages both reservoirs for water supply serving 1.2 million people east of San Francisco Bay (Merz and Chan, 2005). A statistical analysis of the flows from a gaging station downstream of Camanche Dam (USGS Station # 11323500) using the Indicators of Hydrologic Alteration (IHA)

1 software shows the combined effect of both dams on the natural flow regime (Richter *et al.*,
2 1996; Richter *et al.*, 1997) (Fig. 6A). For instance, median flows for the month of May when the
3 highest spring snow-melt flows occur decrease on average from 95 m³/s before Pardee Reservoir
4 to 16 m³/s after Camanche Reservoir (flow records 1905-1929 and 1964-2003, respectively).
5 Since 1964, daily average flows have exceeded the post-dam 10 year return flow interval of 140
6 m³/s in three years: 1986, 1997, and 2006. In addition, the dams have acted as gravel traps,
7 minimizing gravel recruitment downstream of Camanche Reservoir (Pasternack *et al.*, 2004).
8 The flow and sediment budget alterations on the lower Mokelumne River have degraded in-
9 stream habitat and are viewed as main causes for fishery declines (Moyle, 1994). Given the
10 unavailability of flows and gravel, the Federal Energy Regulatory Commission recommended
11 performing gravel replenishment projects to improve fish habitat (Pasternack *et al.*, 2004).

12 River rehabilitation projects to improve habitat for spawning salmon using 8,357 m³ of
13 gravel and cobble have been built in the 1-km reach downstream of Camanche Dam (Wheaton *et*
14 *al.*, 2004b) over the period 1999-2006 and annual placements are on-going. The projects have
15 counteracted channel degradation caused by flow regulation and gravel trapping (Merz *et al.*,
16 2006). Positive ecological effects resulting from the projects include increases in the numbers of
17 fish spawners using the reach, embryo survival to fry stage, macroinvertebrate diversity, and
18 floodplain connectivity (Merz *et al.*, 2004; Merz and Chan, 2005; Elkins *et al.*, 2007).

19 By reducing water depth, increasing water velocity, and changing the morphology of the
20 river bed, gravel augmentation changes the ecological functionality of a site providing
21 appropriate hydraulic conditions to perform ecological functions despite the controlled
22 hydrology of the river (Pasternack, 2008a). The FFM provides a tool to assess the effect of
23 gravel augmentation on the ranges of flows that are functional for theoretical and actual water

1 years, and to assess the changes in the number of days that are functional for specific life stages
2 in a given year. Also, the FFM provides a way to assess how gravel augmentation changes the
3 efficiency of the sites in terms of their capacity to produce functional conditions from the
4 available flows. Performing the analysis in several habitat units is useful to know the overall
5 functionality of restored and un-restored sections of the river and might help explain how gravel
6 augmentation has promoted hydrogeomorphic response and might provide directions on how to
7 proceed in future projects

8 **3.2 Natural Floods on the Yuba River**

9 The largest dam of the North Fork Yuba River is New Bullards Bar Dam (1.2 billion m³),
10 completed in 1970, and the largest dam of the mainstem Yuba River is Englebright Dam (86
11 million m³), built in 1941. The first is a water supply and flood control reservoir, while the latter
12 acts as barrier to block downstream transport of sediment produced during hydraulic mining
13 between 1850 and 1880 (Pasternack, 2008b). IHA analysis of median monthly flows for the
14 mainstem Yuba River (Smartville USGS Station # 11418000) shows a decline in spring-
15 snowmelt flows due to the dams (Fig. 6B). For the month of May, flows decreased on average
16 from 147 m³/s before to 55 m³/s after New Bullards Dam (flow records 1942-1969 and 1970-
17 2003, respectively). The powerhouses at Englebright Dam can only pass 125 m³/s, so discharges
18 greater than that flow overtop the dam. Since 1970, daily average flows have exceeded the post-
19 dam 10-year return flow interval of 2,700 m³/s in three years: 1986, 1997, and 2005. In May and
20 Dec 2005 hourly high flows at Smartville peaked at 1,200 m³/s (7.7 yr return interval) and 3,285
21 m³/s (24 yr return interval) respectively. Despite the sediment trapping effect of Englebright
22 Dam, millions of metric tons of gravel and cobble are stored on the lower Yuba River valley due
23 to hydraulic mining (Pasternack, 2008b).

1 The combined effects of ample sediment present in the river corridor as well as
2 availability of high flows create conditions for the Yuba River to rejuvenate salmon spawning
3 habitat. The particular pool-riffle-run assemblage (500 m long x 250 m wide) located at the apex
4 of Timcubtoo Bend, situated 5 km downstream of Englebright Dam, was repeatedly mapped in
5 1999, 2004, 2005, and 2006. Digital elevation model differencing was used to assess
6 morphological adjustments caused by flows of different magnitudes (Pasternack, 2008b). From
7 1999-2004 there were no major floods, but there was some channel incision focused on riffle
8 crests. From 2004-2005 there were significant areas of both scour and deposition, and the flood
9 accentuated pool-riffle relief by 0.42 m. Looking over a longer time period, aerial photos going
10 back to 1937 demonstrate that this riffle-island complex has persisted for over 70 years, as have
11 six other riffles in Timbuctoo Bend (Pasternack, 2008b).

12 By reworking the coarse hydraulic mining sediment in the river corridor, floods change
13 channel morphology and substrate composition, thus altering local hydraulics. The functional
14 flows analysis of the lower Yuba River provides a tool to analyze the flood-induced changes in
15 1) the ranges of functional flows, 2) the number of days with functional flows for a given water
16 year, and 3) the efficiency of the sites to produce functional conditions. Performing the analysis
17 in several sites is useful in order to observe the spatial distribution of ecological functionality in
18 this dynamic gravel bed river, which is heavily used by spawning fish.

20 **3.3 Selection of Habitat Units for Functional Flows Analysis**

21 Habitat units were selected within river reaches corresponding to riffles that have
22 undergone detailed topographic monitoring before and after events, and from downstream riffles
23 located at non-restored sites on the lower Mokelumne and at a reference site on the lower Yuba

1 River (Fig. 7). On the lower Mokelumne River, sites were selected from riffles located at the
2 furthest upstream reach between Camanche Dam and Mackville Road bridge, which spans 7.8
3 km and corresponds to 32% of total area of the lower Mokelumne River. Habitat units identified
4 for the analysis included three gravel-augmented and three reference riffles (Fig. 7A; 8A). Cross
5 sections were located at each habitat unit. Cross section XS1 was located 237 m downstream of
6 the dam. Initially, this habitat unit was a chute with fast current flowing through two
7 obstructions; after the 1999 gravel addition it became a riffle with depths varying between 0.15
8 and 1.5 m for base flows of 11 m³/s; after the 2004 gravel addition it became a shallower riffle
9 with depths varying between 0.15 and 0.75 m (Fig. 7A). XS2 was located 607 m downstream of
10 the dam. This habitat unit was a degraded deep riffle, and after gravel addition in 2001 it was
11 shaped into a shallow central bar with a downstream riffle (Wheaton *et al.*, 2004b). XS3 was
12 located 295 m downstream of the dam, immediately downstream of XS1, and it became an
13 extension of the XS1 riffle exit after gravel addition in 2005. The section of the river where XS1
14 and XS3 were located presents a steep right bank with encroaching vegetation, and a low-slope
15 left bank with a connected floodplain that has a recreational use. The section where XS2 was
16 located has a vertical right bank formed by a rock outcrop. The reference sites XS4, XS5, and
17 XS6 were located 1,175 m, 1,560 m, and 2,857 m downstream of the dam, respectively, and
18 represent natural riffles that have not been restored (Fig. 7). All the downstream reference sites
19 have a steep left bank and a gently sloping right bank (Fig. 8B).

20 In the Yuba River, habitat units were selected in Timbuctoo Bend: three habitat units
21 were selected at the riffle located at the apex of the bend (Moir and Pasternack, 2008) and a
22 reference habitat unit was selected at the next wide riffle in the downstream direction (Fig. 7B,
23 8C). XS1 and XS2 were located at the riffle entrance and riffle crest respectively, and both

1 eroded during the flood. XS3 was located at the downstream run and accreted during the flood.
2 This section of the river has a connected floodplain, with gravel bars, adjacent channels, small
3 extent of vegetation encroachment, and variable morphology. A main feature of the site was a
4 central bar/island that divides the flow into a main channel to the left and a secondary channel to
5 the right. XS4 was located in a wide, shallow riffle that is heavily used by spawning fish located
6 midway between the apex and the downstream end of Timbuctoo Bend (Fig. 8C).

8 **4 Methods**

9 FFM calculations required ecological, geomorphic, hydrologic, and hydraulic input data.
10 In order to use the algorithm, it was necessary to gather site-specific hydrogeomorphic data of
11 cross section geometry, water surface slope, and grain size distribution. In addition, flow records
12 from USGS stations (#11323500 on the lower Mokelumne River and #11418000 on the lower
13 Yuba River) were used to isolate distinct water year types for both rivers and the event years, or
14 the years pre- and post-gravel augmentation on the lower Mokelumne River and pre- and post-
15 May 2005 flood on the lower Yuba River. The table of functionality, hydrogeomorphic data, and
16 water year types were used as input to perform functional flows calculations.

17 **4.1 Water Year Types**

18 Two types of functional flows analysis were performed: a theoretical analysis using
19 representative water year types for characteristic hydrologic conditions in each river, and an
20 actual analysis using water year data corresponding to the years when events occurred (Table 2).

21 Theoretical water year types were identified using the Flood Regime Characterization
22 (FRC) MATLAB code developed by (Booth *et al.*, 2006)

1 (<http://watershed.ucdavis.edu/pages/programs.html>). The code classifies water year types using
2 mean daily discharge records based on expert input of significant thresholds of flood duration
3 and flood magnitude. One output of the code is the daily flow for each Julian day averaged
4 across all years of the same flood year class.

5 The flow record used for the lower Mokelumne River was 1963-2006 and for the lower
6 Yuba River was 1941-2006. Two water year flood types identified with the FRC code for the
7 lower Mokelumne River included: WY1 that represents a scenario of highly regulated flows with
8 maximum flow of $25 \text{ m}^3/\text{s}$ in the snowmelt season that corresponds to the 1 yr return interval
9 flood; and WY2 that represents a scenario with the highest flows that can be released from the
10 dam with a max flow of $95 \text{ m}^3/\text{s}$ in the rain season that corresponds to the 3 yr return interval
11 flood. Two water year flood types identified with the FRC code for the lower Yuba River
12 included WY1 that represents a scenario of regulated flows with a maximum discharge of 125
13 m^3/s in the snowmelt season and corresponds to the 1 yr return interval flood, and WY2 that
14 represents a scenario with a max flow of $600 \text{ m}^3/\text{s}$ in the rain season and represents a 3 yr return
15 interval flood.

16 Actual water years for the second type of analysis were obtained from daily average flow
17 values from USGS gaging stations. The water years analyzed for the lower Mokelumne River
18 included the pre- and post- gravel augmentation during 1997-2005, and the water years analyzed
19 for the Yuba River included the pre- and post- flood conditions during 2005-2006 (Table 2)
20 (Figs. 9A and 10A; note that WY1 and WY2 were not depicted for limited space).

21 **4.2 Geomorphic Data**

22 Campaigns to collect field data were performed before and after morphologic alterations
23 during the period 1998-2005 on the lower Mokelumne River and during the period 2004-2005 on

1 the Yuba River. Detailed river bed topography data (i.e. 0.5-1.5 pt/m²) was used to build annual
2 channel DEMs using AutoCAD as previously reported (Pasternack *et al.*, 2006; Elkins *et al.*,
3 2007). Cross sections through selected habitat units were sampled from the pre- and post-gravel
4 augmentation surfaces on the lower Mokelumne River, and pre- and post- flood surfaces on the
5 lower Yuba River. The cross section location was centered within the habitat unit of interest to
6 reduce the water constriction effects caused by channel non-uniformity upstream or downstream.
7 A more sophisticated definition of the functional flows model that could be implemented in the
8 future would analyze the spatial sequencing of shear stress to identify functional flows that assist
9 pool-riffle sequence self-maintenance and other dynamics governed by channel non-uniformity
10 (Lisle and Lewis, 1992; MacWilliams *et al.*, 2006). The assumption of uniform flow is
11 appropriate for a channel that does not change cross section geometry in the downstream
12 direction but may need to be checked depending on the conditions of a site (Brown and
13 Pasternack, 2008; Pasternack *et al.*, 2008). For instance, pool-riffle morphologies experience
14 flow convergence, thus requiring the assessment of the relative importance of non-uniform terms
15 to decide if they need to be included in the calculation of shear stress. One way of checking
16 would be to compare the results of steady hydraulic calculations made with a cross-section
17 analyzer against those made with a 1D hydraulic flow model that accounts for backwater
18 conditions (Brown and Pasternack, 2008).

19 Water surface slope as an approximation of river bed slope and grain size distribution for
20 XS1, XS2, and XS3 in both rivers was obtained from previous studies and from unpublished data
21 (Pasternack *et al.*, 2004; Wheaton *et al.*, 2004b; Elkins *et al.*, 2007; Moir and Pasternack, 2008;
22 Pasternack, 2008b) (Table 3). Since water surface slope was reported for a set discharge value, a
23 unique value was obtained for each cross section and was used for the depth-slope product

1 calculations and for stage-discharge geometry relations. This constitutes an assumption, since the
2 water surface slope may change as discharge increases or decreases.

3 In addition to the cross sections sampled from existing DEMs, additional data was
4 collected in November 2005 on both rivers for the reference sites (XS4, XS5, and XS6 in the
5 Mokelumne River, and XS4 on the lower Yuba River). Cross section geometry and bed slope of
6 the reference sites were surveyed with an autolevel, tape, and rod, and the coordinates of the sites
7 were obtained with a Trimble Pathfinder Pro XRS, a real-time kinematic GPS.

8 For habitat units sampled from DEMs and for the reference sites, position and elevation
9 data was used to build stage-discharge relationships using a routine of the functional flows
10 Matlab code. The routine calculates areas for incremental stage values using input cross section
11 geometry and their corresponding velocities using Manning's equation with a typical value of
12 $n=0.043$ for gravel bed rivers (Pasternack *et al.*, 2004) and with their corresponding water
13 surface slopes. The code was used to calculate hydraulic radius and discharge for incremental
14 stage values to obtain the parameters of Eq. 3. Coefficients and exponents are summarized in
15 Table 3.

17 **4.3 Data Analysis**

18 A total of 50 analyses were performed: 30 for the Mokelumne River corresponding to 6
19 cross sections analyzed for 5 distinct water years (i.e. theoretical and actual) and 20 for the Yuba
20 River corresponding to the 5 cross sections analyzed for 4 distinct water years. FFM results were
21 graphed to answer the research questions posited. Results depicted in graphs were grouped first
22 by water year and then by cross section in order to observe trends.

1 To answer research question 1, what are the ranges of flows that are potentially
2 functional?, τ_0^* vs. Q^* curves were graphed indicating the ranges of flows (shaded grey lines)
3 that fell within predetermined bed mobility stages for each water year (Figs. 9B and 10B).
4 Symbols such as circles, squares, or triangles superimposed on the shaded gray lines correspond
5 to the days within a water year that had a functional discharge value that not only fell within
6 specified bed mobility stages but also happened at the time when they were functional for the life
7 stage according to Table 1. Arrows indicate the shift of one cross section from the initial to
8 subsequent locations in the τ_0^* vs. Q^* space due to hydrogeomorphic changes (i.e. gravel
9 augmentation on the lower Mokelumne River and flooding on the lower Yuba River). Functional
10 ranges of Q^* were calculated subtracting the minimum from the maximum functional Q^*
11 occurring in a given water year (Table 4). Cases with zero range of Q^* correspond to absence of
12 ecological functionality, while cases with the highest values of functional ranges of Q^*
13 correspond to a greater opportunity to meet ecological functionality.

14 To answer research question 2, what is the number of days that flows are functional?,
15 counts of the number of days that presented functional flows for each cross section were graphed
16 (Figs. 9C and 10C; Table 4). Higher numbers of days with functional flows correspond to higher
17 ecological functionality performance.

18 To answer research question 3, what is the efficiency of a habitat unit in producing
19 functional flows?, percentage efficiency was estimated as the ratio between functional ranges of
20 Q^* and available ranges of Q^* (Figs. 11 and 12; Table 4). Higher values of efficiency indicate
21 greater opportunity to meet ecological functionality.

22 In order to answer research questions 4-5 it was necessary to perform upstream-
23 downstream and between river comparisons. For these comparisons, results were grouped to

1 evaluate inputs (i.e. slope, median grain size, parameters c and f) and outputs (i.e. ranges of Q^* ,
2 number of days with functional flows, and percent efficiency) of the functional flows model
3 before and after gravel augmentation on the lower Mokelumne River (i.e. 15 before vs. 15 after),
4 before and after May 2005 flood on the lower Yuba River (i.e. 9 before vs. 11 after), among sites
5 within each river (i.e. 24 upstream vs. 6 downstream on the lower Mokelumne River, and 18
6 upstream vs. 2 downstream on the lower Yuba River), and between rivers (i.e. 30 on the lower
7 Mokelumne River vs. 20 on the lower Yuba River) (Table 5). Given that datasets presented
8 differences in standard deviations, they were analyzed with non-parametric statistics. Non-
9 parametric statistics have been widely applied in earth sciences (Pasternack and Brush, 1998).
10 The non-parametric Kolmogorov-Smirnov (KS) two-sample test was performed to determine
11 statistical significance of the difference between data groups to facilitate comparison in research
12 questions to confirm or reject the hypothesis that differences in geomorphic variables caused
13 differences in ecological functionality. This statistical procedure was appropriate to test the
14 relationship between the data groups for the available sample size without making assumptions
15 about the distribution of the data (Statsoft, 1998). The threshold to determine that differences
16 were statistically significant above the 95% confidence level was set at p -level <0.05 (Table 5).

17 Table 5 provides KS results (12 for each comparison group), organized in matrices with 3
18 columns (functional flow outputs) and 4 rows (input variables). First we determined if
19 geomorphic variables (i.e. S , D_{50} , c and f) were in fact statistically different between the two data
20 sets. Then we determined if ecological functionality (i.e. ranges of Q^* , number of days with
21 functional flows, and % efficiency) was different between the two groups. True, or T, was
22 assigned when datasets were statistically different, and False, or F, was assigned when datasets

1 were not statistically different. Then we confirmed or rejected hypothesis following this set of
2 rules: $T \rightarrow T=T$, $T \rightarrow F=F$, $F \rightarrow F=T$, and $F \rightarrow T =F$.

3

4 **5 Functional Flows Analysis Results**

5 Each subsection of the FFM results corresponds to research questions 1 through 5. In
6 addition to description of graphs and presentation of calculation outputs, each subsection refers
7 to the KS test reported in Table 5 for confirmation or rejection of the overall hypothesis.

8 **5.1 Change in functional ranges of Q^***

9 For the lower Mokelumne River, results indicated that river rehabilitation caused a
10 vertical shift in τ_0^* vs. Q^* curves for XS1 and XS3 from a non-functional domain (i.e. SB) to a
11 partially functional domain (i.e. IFM or FM) for all water years (Figs. 9B). The range of
12 functional flows was increased at XS1 and XS3 for all water years and at XS2 for WY1. Note
13 that in this case the increase in range of functional flows does not imply achieving full
14 functionality because the ranges of functional flows (i.e. falling in either IFM or FM but not in
15 both) did not support all of the essential functions (i.e. missing IFM or FM). The effect of gravel
16 augmentation on geomorphic variables S , D_{50} , c , and f was statistically significant above the
17 95% confidence level ($p < 0.01$, < 0.05 , < 0.01 , < 0.005 respectively), but was not statistically
18 significant ($p < 0.10$) for functional range of Q^* . The hypothesis was rejected for all 4
19 combinations of functional ranges of Q^* with hydrogeomorphic variables (Table 5A).

20 For the lower Yuba River, results indicated that natural floods caused a lateral shift in τ_0^*
21 vs. Q^* curves for XS1 and XS2 maintaining them within a functional domain. τ_0^* vs. Q^* curves
22 for XS3 presented a diagonal shift (Figs. 10B). Lateral shifts to the right and a diagonal shift

1 increased the ranges of functional flows for all cross sections for all water years. Changes before
2 and after the flood on hydrogeomorphic variables S , D_{50} , c , and f and on the functional ranges of
3 Q^* was statistically significant above the 97.5 confidence level ($p < 0.001$, < 0.001 , < 0.025 ,
4 < 0.0025 , < 0.005 respectively). The hypothesis was accepted for all 4 combinations of functional
5 ranges of Q^* and geomorphic variables (Table 5A).

6 **5.2 Change in the number of days that are functional**

7 For the lower Mokelumne River, results indicated that river rehabilitation increased the
8 number of days with functional flows for XS1 and XS3 for all water years, while it increased the
9 number of days with functional flows for XS2 for WY2 only (Figs. 9C). On the lower
10 Mokelumne River, the mean value of days with functional flows was 60 before and 138 after
11 gravel augmentation. The effect of gravel augmentation on the number of days with functional
12 flows was statistically significant above the 95% confidence level ($p < 0.05$). The hypothesis was
13 accepted for all 4 combinations of number of days with functional flows and geomorphic
14 variables (Table 5A).

15 For the lower Yuba River, results indicated the May 2005 natural flood maintained the
16 number of days with functional flows for WY1 and WY2 while it increased the number of days
17 with functional flows for the year when the event happened for XS1 and XS2 (Figs. 10C). Also,
18 the flood increased the number of days with functional flows for all 4 water years for XS3. Even
19 though the mean value of days with functional flows across all 4 water years was 157 before and
20 174 after the May 2005 flood, the change in the number of days with functional flows was not
21 statistically significant ($p > 0.10$). The hypothesis was rejected for all 4 combinations of number
22 of days with functional flows and geomorphic variables (Table 5A).

5.3 Change in efficiency of a habitat unit to produce functional flows

For the lower Mokelumne River, results indicated that after gravel augmentation XS1 and XS3 increased their efficiency to produce functional flows for all water years, while XS2 increased its efficiency to produce functional flows for WY2 only. The change in efficiency was not statistically significant ($p < 0.10$). The hypothesis was rejected for all 4 combinations of % efficiency and geomorphic variables (Table 5A).

For the lower Yuba River, results indicated that the May 2005 natural flood maintained the efficiency to produce functional flows for XS1 and XS2 for WY1 and WY2, reduced the efficiency for XS3 for WY2, and increased the efficiency for all cross sections for the actual WY. The change in % efficiency was not statistically significant ($p > 0.10$). The hypothesis was rejected for all 4 combinations of % efficiency and geomorphic variables (Table 5A).

5.4 Overall functionality of a river reach

The overall functionality of each river reach was analyzed by comparing the functionality of habitat units located at detailed monitoring and reference sites within the same river. For the Mokelumne River, the differences in the values of c and f between upstream and reference sites were statistically significant ($p < 0.001$ and $p < 0.05$ respectively), while the differences in values of S , D_{50} , and functional flows results were not statistically significant ($p > 0.10$). Reference sites of the lower Mokelumne River presented some degree of functionality measured from the occurrence of days with functional flows and from the efficiency to produce functional flows. Curves of τ_0^* vs. Q^* for XS4 and XS6 were within the functional domain for WY1 and WY2, while the curve of τ_0^* vs. Q^* for XS5 presented a small section within the functional domain for WY2 only. XS4 and XS6 presented days with functional flows for both water years, while XS5

1 presented days with functional flows for WY2 only. XS4 and XS6 presented above average
2 efficiency for WY1 and WY2, while XS5 presents above average efficiency for WY2 only
3 (Table 4). The hypothesis was confirmed for 6 and rejected for 6 combinations of functional
4 flows outputs and hydrogeomorphic variables inputs (Table 5B).

5 For the lower Yuba River, hydrogeomorphic variables and functional flows outputs of
6 upstream and downstream sites were statistically similar ($p > 0.05$). XS4 located at the reference
7 site presented lower functionality than upstream reaches for WY1 as shown by its lower section
8 of the τ_0^* vs. Q^* curve within the functional domain and by the low number of days with
9 functional flows. In contrast, XS4 presented functionality comparable to upstream reaches for
10 WY2 represented by its high section of the τ_0^* vs. Q^* curve within the functional domain and by
11 its high number of days with functional flows. XS4 presented lower than average efficiency for
12 WY1 and higher than average efficiency for WY2 (Table 4). The hypothesis was confirmed for
13 all 12 combinations of functional flows outputs and hydrogeomorphic variables inputs (Table
14 5B).

15 **5.5 Comparison among rivers**

16 The lower Yuba River presented better flow functionality than the lower Mokelumne
17 River determined by the location of τ_0^* vs. Q^* curves for all cross sections within the functional
18 domains and by a higher number of days with functional flows. In addition, the efficiency of the
19 lower Yuba River sites was higher on average than that of the lower Mokelumne River sites.
20 Slope, D_{50} , c , and functional flows outputs were statistically different ($p < 0.001$) while f was
21 statically similar ($p > 0.10$). The hypothesis was accepted for 9 combinations of functional flows
22 outputs and hydrogeomorphic variables inputs (Table 5C).

23

6 Discussion

The analysis of ecological functionality using functional flows reflects how changes of geomorphic variables due to hydrogeomorphic processes modify the suitability of in-stream physical habitat for fall-run Chinook Salmon. For each habitat unit, ranges of flows, number of days that flows are functional, and efficiency to produce functional flows were obtained from Eqs. 5 and 6. Rewriting Eq. 5 in terms of Q^* yields:

$$\tau_0^* = \frac{\rho}{(\rho_s - \rho)} \frac{S}{D_{50}} c (g^{1/2} D_{50}^{5/2} Q^*)^f$$

where each input variable influences the final result depending on its effect on the value of τ_0^* .

The exponent f determines the slope of the curve τ_0^* vs. Q^* , the D_{50} is related non-linearly to

τ_0^* , and S and c are related linearly to τ_0^* . Overall, within the ranges found at the cross sections

studied, lower values of f promoted lower depth response to discharge increments that may be

beneficial for spawning habitat (i.e. such response is found in shallow riffles), large values of D_{50}

promoted higher upper thresholds of Q^* increasing the ranges of functional flows, and higher

values of c and S promoted the vertical shift up of τ_0^* vs. Q^* curves from stable bed to

functional transport regimes (i.e. superficial fines mobility, interstitial fines mobility, and full

mobility).

The next subsections include a discussion of results obtained analyzing the effect of

geomorphic variables on functional flows output exclusively. Although measured increases in

the numbers of fish spawners using the reach, embryo survival to fry stage, macroinvertebrate

diversity, and floodplain connectivity are metrics of ecological improvement, they are not

comparable to ranges of functional flows, number of days with functional flows and percent

efficiency. In order to use field data to compare to FFM results, it would be necessary to perform

1 field campaigns to measure whether or not each day's flow that was classified as functional was
2 in fact functional by testing whether or not fish used the habitat for a specific ecologic function.

3

4 **6.1 Effect of geomorphic changes on functional ranges of Q^***

5 The rejection of the hypothesis that changes in geomorphic variables due to gravel
6 augmentation modified the ecological response for bed occupation and bed preparation as
7 measured by the difference in functional ranges of flows on the lower Mokelumne River
8 indicates that the alteration of the morphologic variables S , D_{50} , c and f did not cause an
9 ecological improvement of the habitat. The positive change of S into higher values that caused a
10 shift of the τ_o^* vs. Q^* curves from a stable bed into functional transport regimes together with
11 the positive change of the variable f into lower values that signify a lower depth response to
12 increments in discharge increased the functional ranges of flows to some level. The positive
13 effects of S and f were counteracted by the negative change of D_{50} into a larger value that
14 reduced the upper threshold of the ranges of Q observed in the lateral shift to the left of the τ_o^*
15 vs. Q^* curves. The manipulation of the morphology of the channel alone was not sufficient to
16 create a statistically significant difference after gravel augmentation of the functionality of the
17 habitat as measured by the increase of functional ranges of flows and cannot substitute the need
18 for larger flows that would increase the ranges of functional flows to a level that is statistically
19 significant.

20 The confirmation of the hypothesis that changes in geomorphic variables due to the May
21 2005 flood modified the ecological response as measured by the difference in functional ranges
22 of flows on the lower Yuba River indicates that the changes in the morphologic variables S , D_{50} ,
23 c and f were sufficient to improve the ecological conditions of the habitat. Despite the fact that

1 the variable f increased changing into values that are theoretically less functional promoting
2 greater depth increments to increments in discharge that may be negative to the habitat, the
3 combined effect of a lower slope, smaller grain size, and available flows provided the conditions
4 to increase the functional ranges of flows.

5 **6.2 Effect of geomorphic changes on number of days with functional flows**

6 The confirmation of the hypothesis that changes in geomorphic variables due to gravel
7 augmentation modified the ecological response as measured by the difference in the number of
8 days with functional flows on the lower Mokelumne River suggests that achievement of a lower
9 depth response to increments in discharge by reducing the values of f and the achievement of a
10 functional sediment transport stage by increasing the values of S allowed the available low flows
11 to provide functional habitat conditions during some of the crucial spawning life stages.

12 The rejection of the hypothesis that changes in geomorphic variables due the May 2005
13 modified the ecological response as measured by the difference in the number of days with
14 functional flows on the lower Yuba River indicates that the geomorphic improvements after the
15 flood were not sufficient to improve the number of days with functional flows. An explanation of
16 this result is that sites presented a high count of days with functional flows even before the flood,
17 and the geomorphic changes after the flood did not impact the results significantly. This suggests
18 that there may be a threshold of number of days with functional flows for sites, depending on
19 hydrologic conditions, above which it is unlikely to increase.

20 **6.3 Effect of geomorphic changes on the efficiency to produce functional** 21 **flows**

22 The efficiency of a habitat unit to produce functional flows is a metric that combines the

1 hydrologic and geomorphic response with the ecological requirements for life stages. The
2 complex, non-linear interaction among variables yields a variety of hydrogeomorphic responses
3 and the consequent variability in efficiency to produce functional flows. The available ranges of
4 flows may fall within a non-functional domain causing the absence of functional ranges of flows
5 or the available ranges of flow may fall within a functional domain in which case the presence of
6 functional ranges of flows depends on the time series of flows that may or may not produce
7 functional sediment transport regimes at the appropriate times for each life stage.

8 Despite the given conditions to increase efficiency at both rivers due to the increase of
9 number of days with functional flows on the lower Mokelumne River and due to the increase of
10 functional ranges of flows on the lower Yuba River, the hypothesis that changes in geomorphic
11 variables modified the ecological response as measured by % efficiency was rejected, indicating
12 that this metric did not reflect the habitat improvement shown by the other two functional flows
13 outputs.

14 **6.4 Functional flows analysis at the reach scale**

15 In the context of comparing cross sections within the same reach, geomorphic similarity
16 can be defined for cross sections with geomorphic variables that are statistically similar (i.e. the
17 results of the KS statistical test are negative). Likewise, the similarity of the ecological
18 performance can be defined for cross sections with functional flows outputs that are statistically
19 similar. The confirmation of the hypothesis that similarity of the geomorphic variables S and D_{50}
20 caused similarity of ecological performance as measured by functional ranges of flows, number
21 of days with functional flows, and efficiency between restored and un-restored sites within the
22 same river reach on the lower Mokelumne River indicates that un-restored sites may not need the
23 same level of channel morphology modification that restored sites underwent. The rejection of

1 the hypothesis for the geomorphic variables c and f indicates that un-restored sites may need
2 channel geometry improvements such as reduction of f values to decrease depth response to
3 discharge increments and decrease incision that generated positive effects on the restored sites.

4 The confirmation of the hypothesis that similarity of all the geomorphic variables analyzed
5 caused similarity of ecological performance as measured by functional ranges of flows, number
6 of days with functional flows, and efficiency between pre and post-flood sites within the same
7 river reach at the Yuba River indicates that the larger scale processes that control the
8 geomorphology of the reach also control ecological performance of the habitat in this section of
9 the river.

10 **6.5 Functional flows analysis at the regional scale**

11 In the context of comparing rivers within the same region such as the Sierra Nevada,
12 statistical differences of geomorphic variables and ecological performance reflect the importance
13 of the geomorphology and history of each watershed. The confirmation of the hypothesis that
14 differences of geomorphic variables S , D_{50} , and c cause differences of ecological performance as
15 measured by the functional ranges of flows, number of days with functional flows, and efficiency
16 between the lower Mokelumne River and the lower Yuba River indicates that processes at each
17 watershed, such as geomorphic controls and local human impacts, rather than regional processes,
18 such as climate, control the ecological performance of each river. Higher values of S and higher
19 values of D_{50} have the potential to cause lower functionality on the lower Yuba River with
20 respect to the lower Mokelumne River by reducing the span of τ_o^* vs. Q^* curves within
21 functional sediment transport domains. However, values of functional ranges of Q^* , number of
22 days with functional flows, and % efficiency are larger on the lower Yuba River than on the
23 lower Mokelumne River. Lower values of c may contribute partially to the effect of greater

1 functionality on the lower Yuba River because they cause a higher span within functional
2 sediment transport stages. Yet the largest factor contributing to better functional flows outputs on
3 the lower Yuba River is larger flow availability, which promotes ecological functionality despite
4 the overall lower geomorphic performance of the lower Yuba River with respect to the lower
5 Mokelumne River.

6 **6.6 Key Lessons**

7 Gravel augmentation on the lower Mokelumne River between 1999-2006 has increased the
8 number of days with functional flows for fall-run Chinook salmon for the study sites but has not
9 impacted significantly the functional ranges of flows. Hydrogeomorphic variables have reached a
10 functional stage after gravel augmentation as observed by the location of τ_o^* vs Q^* curves in
11 functional domains. The next possible stage to increase ecological functionality in the restored
12 sites is to increase available ranges of flows at the appropriate times during the year in order to
13 increase functional ranges of flows.

14 The May 2005 flood on the lower Yuba River increased functional ranges of flows by
15 shifting τ_o^* vs Q^* laterally. Although the May 2005 flood increased the number of days with
16 functional flows at the study sites, there was a high occurrence of days with functional flows
17 before the flood and the consequent effect of the flood on this analysis output is insignificant.

18 The metric of percentage efficiency did not reflect the ecological improvements of number
19 of days with functional flows on the lower Mokelumne River and of functional ranges of Q^* on
20 the lower Yuba River. This is a complex metric that requires several steps for calculation and
21 involves several variables that may counteract each other. According to the results of this study,
22 the work invested in obtaining this metric does not provide additional information that is helpful
23 for understanding ecological functionality in rivers.

1 Detailed monitoring sites and reference sites present similar ecological functionality in
2 both rivers. Despite the local effects of gravel augmentation that have changed local geometry on
3 the lower Mokelumne River sites, ecological functionality of reference sites is similar to that of
4 restored sites indicating that reference sites may not need abrupt gravel augmentation projects to
5 improve their habitat. In the Yuba River, study cross-sections at the apex of Timbuctoo Bend and
6 the reference site present similar ecological functionality, indicating the uniformity of conditions
7 within the reach to provide habitat quality for fall-run Chinook salmon.

8 The lower Mokelumne and Yuba Rivers in general present differences in ecological
9 functionality. Overall, the lower Mokelumne River has a characteristic geomorphic functionality
10 that comes from the combination of hydrogeomorphic variables such as slope, grain size
11 distribution, and cross section geometry that produce ecological functionality despite low flow
12 availability. On the other hand, the lower Yuba River also presents geomorphic functionality that
13 is complemented by a hydrologic functionality that comes from ample flow availability for an
14 optimal combination of hydrologic and geomorphologic conditions for ecological functionality.

16 **7 Conclusions**

17 Differences in local geometry, within a river reach and between rivers, may affect how
18 habitat units respond to functional flows metrics such as number of days with functional flows or
19 ranges of functional flows. Such hydrogeomorphic and ecologic functionality differences are
20 governed by site specific conditions and processes. Rapid geomorphic changes are one type of
21 river processes that have the capacity to alter hydraulics that in turn affect sediment transport
22 stages and ecological response of the river bed. Depending on the direction of the morphologic
23 change, such alterations may be positive for the physical habitat. When geomorphic change

1 promotes the proper combination of geomorphic variables and hydrology it also induces the
2 conditions for improved ecological functionality. Sites with suitable combinations of slope, grain
3 size distribution, and cross section geometry may have the potential to create ecological
4 functionality. However, ecological functionality will only be provided if there is ample
5 availability of functional ranges of flows working with local morphology and hydraulics to
6 provide sediment transport stages that are functional for fall-run Chinook salmon life stages. The
7 application of the functional flows analysis presented in this paper contributes to the current
8 knowledge of interactions between hydraulics, geomorphology, and ecology indicating the
9 pertinence of this approach to the crucial understanding of the effects of physical processes on
10 ecological response.

11

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Figures Captions

Figure 1. Interaction of hydrogeomorphic processes that control physical habitat conditions for ecological functionality.

Examples of each process or function are given inside each circle. Convulsive events are in italics.

Figure 2. Location of Mokelumne River and Yuba River study reaches

Figure 3. Life stages of Fall-run Chinook salmon in relation to flow magnitude and timing

A) Bed occupation and bed preparation ecological functions timing for fall-run Chinook salmon freshwater life stage; B) Water year flow magnitudes at the Yuba River, CA and examples of functional flows for the ecological functions in A).

Figure 4. Functional flows classification for fall-run Chinook salmon ecological functions

Non-dimensional shear stress time series for riffle cross section on the lower Yuba River with $S=0.046$ and $D_{50}=0.068$ for the water year depicted in 3B) with functional (solid line) and non-functional (dashed line) transport regimes according to Table 1. After day 330 BO stands for Bed Occupation, N-f stands for non-functional, and F stands for functional.

Figure 5. τ_* vs Q^* curve for example in Figure 4

Non-dimensional shear stress vs non-dimensional discharge for identification of functional ranges of flows using same example in Figs. 3 and 4. Available Q^* (gray line) refers to ranges of flows within a water year that fall within specified bed mobility stages, Functional Q^* (triangle symbol) refer to ranges of flow within a water year that fall within specified bed mobility stages and happen at the time when they are functional for the life stage. FM stands for full mobility, IFM stands for interstitial fines mobility, SFM stands for superficial fines mobility, and SB stands for stable bed

Figure 6. IHA-RVA analysis of flow records

1 A) Mokelumne River USGS Station # 11323500, B) Yuba River USGS Station # 11418000.
 2 RVA refers to “range of variability approach” targets defined by Richter et. al. 1997. Middle
 3 RVA low and middle RVA high correspond to 25th and 75th percentile levels of the monthly
 4 average flows.

5
 6 **Figure 7. Habitat units for functional flows analysis**

7 Projected coordinate system: NAD_1983_StatePlane_California_III for Mokelumne River, and
 8 NAD_1983_StatePlane_California_IV for Yuba River,. Coordinates in meters for A) Mokelumne
 9 River sites XS1 (1,953,735, 691,786), XS2 (1,953,739, 691,672), XS3 (1,953,383, 691,716),
 10 XS4 (1,953,435, 691,661), XS5 (1,953,697, 691,777), XS6 (1,953,696, 691,690); and B) Yuba
 11 River sites XS1 (2,059,431, 674,116), XS2 (2,059,388, 674,141), XS3 (2,059,179, 674,222),
 12 XS4 (1,928,546, 803,447)

13 **Figure 8. Cross sections geometry**

14 A) Mokelumne River cross sections at restored riffles before (dashed line) and after (solid line)
 15 gravel augmentation, B) Mokelumne River cross sections at reference riffles, C) Yuba River
 16 cross sections before (dashed line) and after (solid line) May 20005 natural flood and reference
 17 site (XS4).

18 **Figure 9. Functional flows analysis of restored riffles for Actual WY at Mokelumne River**

19 A) Julian water year discharge time series for water years before and after gravel augmentation,
 20 B) $\tau_{.}^*$ vs Q^* curves - gray lines correspond to the actual ranges of Q^* produced by the water year
 21 and symbols correspond to the days with functional ranges of Q^* that produced a functional
 22 sediment transport stage for a specific ecological function according to the table of functionality
 23 (Table 1), symbols are not depicted when days with functional ranges of Q^* are absent for the

1 specific site and ecological function, C) Count of number of days with functional flows for water
2 year in A)

3 **Figure 10. Functional flows analysis of all sites for Actual WY at Yuba River**

4 A) Julian water year discharge time series for actual water years before and after May 2005
5 flood, B) and C) captions are the same as Fig. 9. See insert in Figure 10B for a detail of the
6 curves for year 04-05 for all three cross sections.

7 **Figure 11. Efficiency of habitat units to produce functional flows for Mokelumne River**

8 Empty circles indicate minimum and maximum available Q^* for the water year, and solid circles
9 indicate minimum and maximum functional Q^* for the water year for A) Restored sites, B)
10 Reference sites

11 **Figure 12. Efficiency of habitat units to produce functional flows for Yuba River**

12 Empty circles indicate minimum and maximum available Q^* for the water year, and solid circles
13 indicate minimum and maximum functional Q^* for the water year.

14

15 **Tables Captions**

16 **Table 1. Table of functionality.**

17 Flow magnitude and bed mobility stages delimited by Shields stress are used to determine
18 functionality for bed occupation and bed preparation ecological functions during the spawning
19 life stage. “Functional” refers to flow magnitudes associated with bed mobility stages that favor
20 the life stage. “Non-functional” refers to flow magnitudes associated with bed mobility stages
21 that hinder the life stage.

22 **Table 2. Summary of functional flows analysis**

1 Type of analysis is theoretical for WY1 and WY2, and Actual for water years when the events
2 occurred. Timeline of events represent water years (horizontal arrows) and the approximate date
3 of the events occurrence (vertical arrows). Sites marked with * are reference sites. Sites marked
4 with ** were analyzed twice for WY1 and WY2 due to the occurrence of two different gravel
5 augmentation projects at the same site. A total of 30 analysis were performed on the lower
6 Mokelumne River, and a total of 20 analysis were performed on the lower Yuba River.

7 **Table 3. Summary of physical parameters for functional flows analysis**

8 Cross sections geometry, slope, and median grain size were obtained from data reported in
9 previous studies as indicated next to each value and from data collected for this study. Data
10 sources are (1) Merz et al 2005 , (2) Elkins et al 2007 , (3) Wheaton 2003, (4) This Study, (5)
11 Moir and Pasternack Submitted. Parameters c and f were obtained from cross section geometry
12 relations developed for each cross section geometry.

13 **Table 4. Summary of functional flows analysis comparison criteria and outputs**

14 Comparison criteria are before/after gravel augmentation on the lower Mokelumne River or
15 flood on the lower Yuba River (B/A), and detailed monitoring sites or reference site (D/R).
16 Functional flows outputs are available ranges of Q^* , functional ranges of Q^* , # days with
17 functional flows, and % efficiency ($100 \times \text{functional ranges of } Q^* / \text{available ranges of } Q^*$).

18 **Table 4. Summary of functional flows analysis comparison criteria and outputs**

19 **(Continuation)**

20 **Table 5. Hypothesis testing and statistical significance of comparisons among geomorphic**
21 **input variables and functional flows outputs datasets.**

22 Differences between datasets were considered statistically significant for $p\text{-level} < 0.05$. Table
23 contains $p\text{-level}$ and Kolmogorov-Smirnov (KS) test results for each dataset comparison. A)

1 Before and after rapid alteration of channel morphology, B) Detailed monitoring sites vs.
2 reference sites, C) Mokelumne River vs. Yuba River. rQ^* stands for functional ranges of Q^* ,
3 #DFE stands for # days with functional flows, %Eff stands for % efficiency. Hypothesis:
4 statistically significant differences in hydrogeomorphic conditions cause statistically significant
5 differences in ecological performance of the physical habitat. Hypothesis confirmation or
6 rejection is indicated at the crossing of inputs (left column) vs. outputs (top row) for datasets
7 compared. Confirmation of the hypothesis is provided according to material conditional truth
8 rules $T \rightarrow T = T$, $T \rightarrow F = F$, $F \rightarrow F = T$. The combination $F \rightarrow T$ was considered F.

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1	2A	2B	3	4	5	6
Flow Magnitude/ Bed Mobility Stage	Functional Flows Delimiters			Bed Occupation		
	In terms of τ_0 vs τ_{c50}	In terms of τ_0^* vs Shields #	Spawning	Embryo Incubation	Emergency	Preparation
High/ Full mobility (FM)	$\tau_0 > \tau_{c50}$	$0.06 < \tau_0^* < 0.1$	Non-functional	Non-functional	Non-functional	Functional
Intermediate High/ Interstitial fines mobility (IFM)	$\tau_0 = \tau_{c50}$	$0.03 < \tau_0^* < 0.06$	Non-functional	Non-functional	Non-functional	Functional
Intermediate Low/ Superficial fines mobility (SFM)	$\tau_0 < \tau_{c50}$	$0.01 < \tau_0^* < 0.03$	Functional	Functional	Functional	Non-functional
Low/ Stable bed (SB)	$\tau_0 \ll \tau_{c50}$	$\tau_0^* < 0.01$	Functional	Non-functional	Functional	Non-functional

River	XS Name	Date Surface/ Surveyed	S	D ₅₀	c	f
Mokelumne River	XS1	97-98	0.0001 ⁽¹⁾	40.0 ⁽¹⁾	0.38	0.45
		98-99	0.0038 ⁽¹⁾	50.0 ⁽¹⁾	0.22	0.39
		02-03	0.0020 ⁽²⁾	50.4 ⁽²⁾	0.18	0.43
		03-04	0.0080 ⁽²⁾	50.4 ⁽²⁾	0.17	0.39
	XS2	99-00	0.0003 ⁽³⁾	68.0 ⁽³⁾	0.35	0.41
		00-01	0.0006 ⁽³⁾	55.0 ⁽³⁾	0.30	0.39
	XS3	03-04	0.0003 ⁽¹⁾	47.5 ⁽¹⁾	0.34	0.46
		04-05	0.0018 ⁽⁴⁾	71.0 ⁽⁴⁾	0.22	0.39
	XS4	Nov 2005	0.0034 ⁽⁴⁾	67.0 ⁽⁴⁾	0.12	0.52
	XS5	Nov 2005	0.0012 ⁽⁴⁾	53.0 ⁽⁴⁾	0.16	0.38
XS6	Nov 2005	0.0052 ⁽⁴⁾	40.0 ⁽⁴⁾	0.11	0.51	
Yuba River	XS1	03-04	0.0069 ⁽⁵⁾	101.1 ⁽⁵⁾	0.16	0.37
		04-05	0.0046 ⁽⁴⁾	60.7 ⁽⁴⁾	0.09	0.47
	XS2	03-04	0.0069 ⁽⁵⁾	101.1 ⁽⁵⁾	0.18	0.35
		04-05	0.0046 ⁽⁴⁾	78.0 ⁽⁴⁾	0.15	0.39
	XS3	03-04	0.0046 ⁽⁵⁾	179.8 ⁽⁵⁾	0.13	0.41
		04-05	0.0039 ⁽⁴⁾	69.3 ⁽⁴⁾	0.14	0.42
	XS4	Nov 2005	0.0011 ⁽⁴⁾	66.0 ⁽⁴⁾	0.10	0.50

Comparison							Functional Flows Analysis			
River	Water Year Type	XS	Site	Before/After	Detailed monitoring site/Reference	Available Range Q*	Functional Range Q*	# Days with functional flows	% Efficiency	
Mokelumne River	WY1		97-98	B	D	10,948	-	0	0	
			98-99	A	D	10,948	10,948	248	100	
			02-03	A	D	10,733	9,715	179	91	
			03-04	A	D	10,733	8,176	185	76	
			99-00	B	D	5,076	-	0	0	
			00-01	A	D	8,627	-	0	0	
	WY2			03-04	B	D	12,446	-	0	0
				04-05	A	D	4,557	2,290	75	52
				Nov05	B	R	5,267	4,768	179	91
				Nov05	B	R	9,464	-	0	0
				Nov05	B	R	19,126	19,126	250	100
				97-98	B	D	81,580	-	0	0
	Actual WY			98-99	A	D	46,699	31,461	190	67
				02-03	A	D	45,778	38,363	174	84
				03-04	A	D	45,778	25,687	185	56
				99-00	B	D	21,650	-	0	0
				00-01	A	D	36,798	16,450	50	45
				03-04	B	D	53,089	10,200	14	19
	Actual WY			04-05	A	D	19,435	18,853	174	97
				Nov05	B	R	22,467	15,136	200	67
				Nov05	B	R	40,369	20,605	73	51
				Nov05	B	R	81,580	45,777	185	56
				97-98	B	D	95,006	-	0	0
				98-99	A	D	45,064	15,129	232	34
Actual WY			02-03	A	D	27,742	1,729	179	6	
			03-04	A	D	35,626	23,809	155	67	
			99-00	B	D	15,559	-	0	0	
			00-01	A	D	3,608	-	0	0	
			03-04	B	D	23,785	-	0	0	
			04-05	A	D	15,125	5,600	43	36	

Comparison Criteria				Functional Flows Analysis Outputs					
River	Water Year Type	XS	Site	Before/After	Detailed monitoring site/Reference site	Available Range Q*	Functional Range Q*	# Days with functional flows	% Efficiency
Yuba River	WY1	XS1	03-04	B	D	10,160	10,054	184	99
			04-05	A	D	36,405	36,024	190	99
		XS2	03-04	B	D	10,160	10,054	179	99
			04-05	A	D	19,458	19,254	162	99
	XS3	03-04	B	D	2,409	1,219	86	51	
		04-05	A	D	26,099	25,826	160	99	
	XS4	Nov05	A	R	29,506	2,924	7	10	
	WY2	XS1	03-04	B	D	58,405	21,645	177	37
			04-05	A	D	209,272	77,559	176	37
		XS2	03-04	B	D	58,405	21,645	177	37
			04-05	A	D	111,852	41,454	170	37
		XS3	03-04	B	D	13,847	13,136	112	95
			04-05	A	D	150,030	55,603	173	37
		XS4	Nov05	A	R	169,615	148,392	101	87
		Actual WY	XS1	03-04	B	D	10,944	4,954	182
	04-05			A	D	306,674	199,965	265	65
XS2	03-04		B	D	10,944	4,954	216	45	
	04-05		A	D	163,912	163,912	254	100	
XS3	03-04	B	D	2,595	1,874	99	72		
	04-05	A	D	219,859	219,859	255	100		

A. Mokelumne River (n=15 before vs. n=15 after)

p-level	rQ*		#DFF		%Eff	
	p < .10	p < .05	T	F	p < .10	F
	KS test					
S	p < .01		F	T	F	F
D ₅₀	p < .05		F	T	F	F
c	p < .01		F	T	F	F
f	p < .005		F	T	F	F

Yuba River (n=15 before vs. n=15 after)

p-level	rQ*		#DFF		%Eff	
	p < .005	p > .10	T	F	p > .10	F
	KS test					
S	p < .001		T	F	F	F
D ₅₀	p < .001		T	F	F	F
c	p < .025		T	F	F	F
f	p < .025		T	F	F	F

B. Mokelumne River (n=24 restored vs. n=6 reference)

p-level	rQ*		#DFF		%Eff	
	p > .10	p > .10	F	F	p > .10	F
	KS test					
S	p > .10		T	T	T	T
D ₅₀	p > .10		T	T	T	T
c	p < .001		F	F	F	F
f	p < .05		F	F	F	F

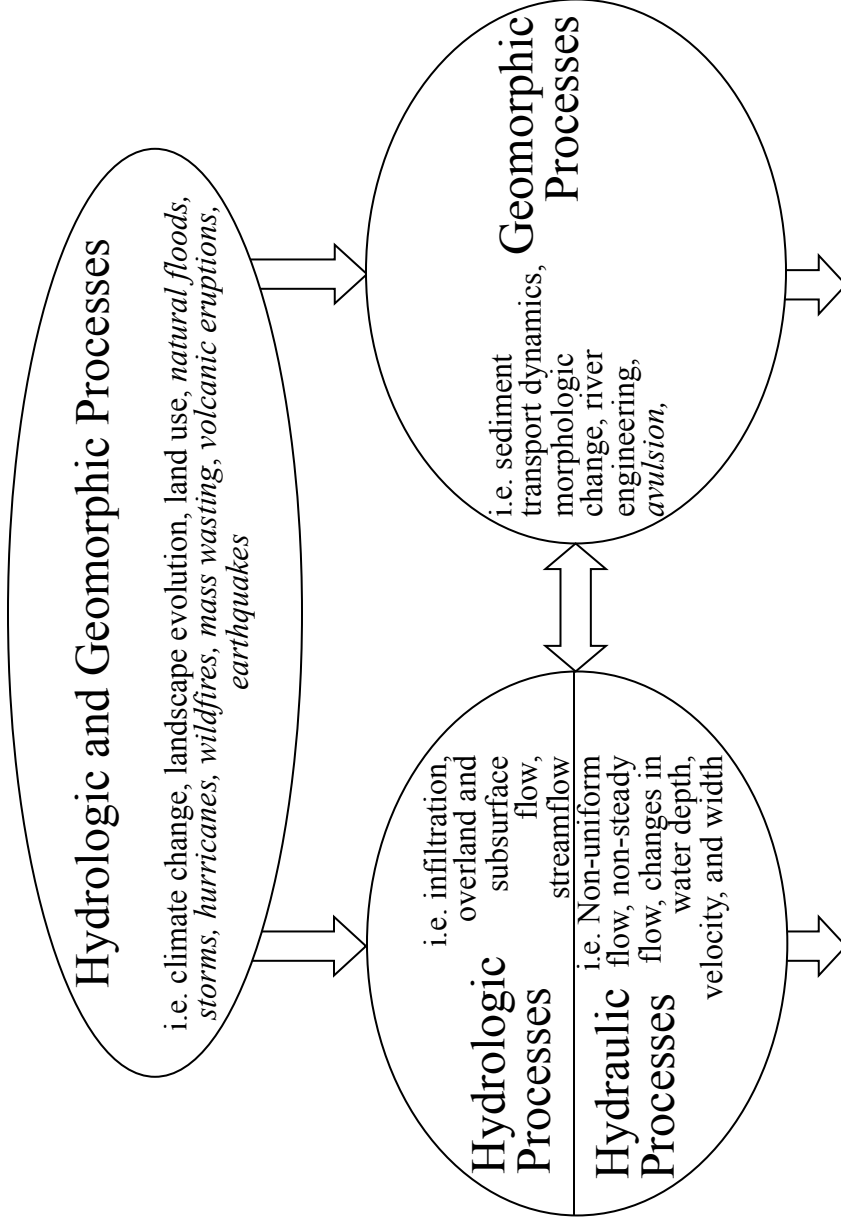
Yuba River (n=18 apex vs. 2 reference)

p-level	rQ*		#DFF		%Eff	
	p > .10	p > .10	F	F	p > .10	F
	KS test					
S	p < .10		F	T	T	T
D ₅₀	p > .10		F	T	T	T
c	p > .10		F	T	T	T
f	p < .10		F	T	T	T

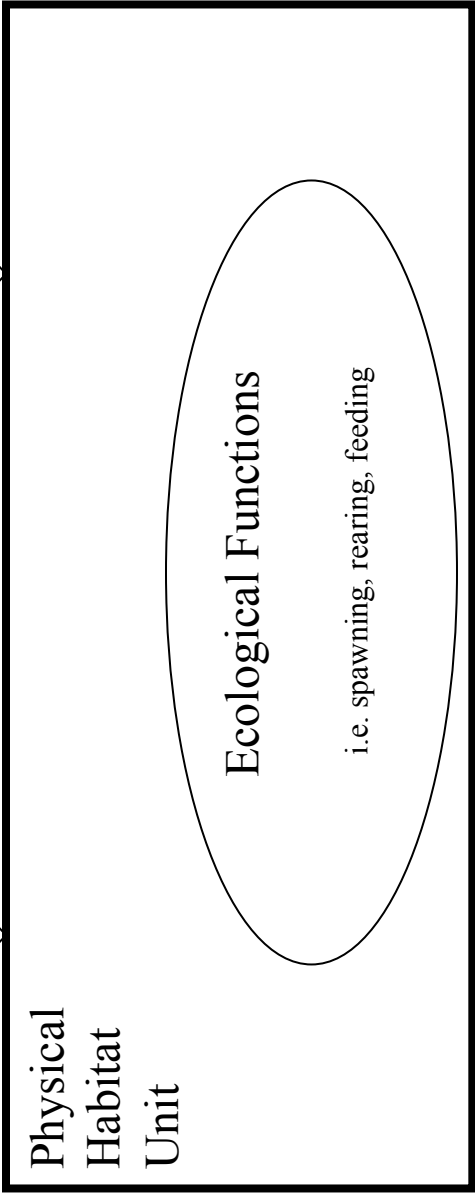
C. Mokelumne River (n=30) vs. Yuba River (n=20)

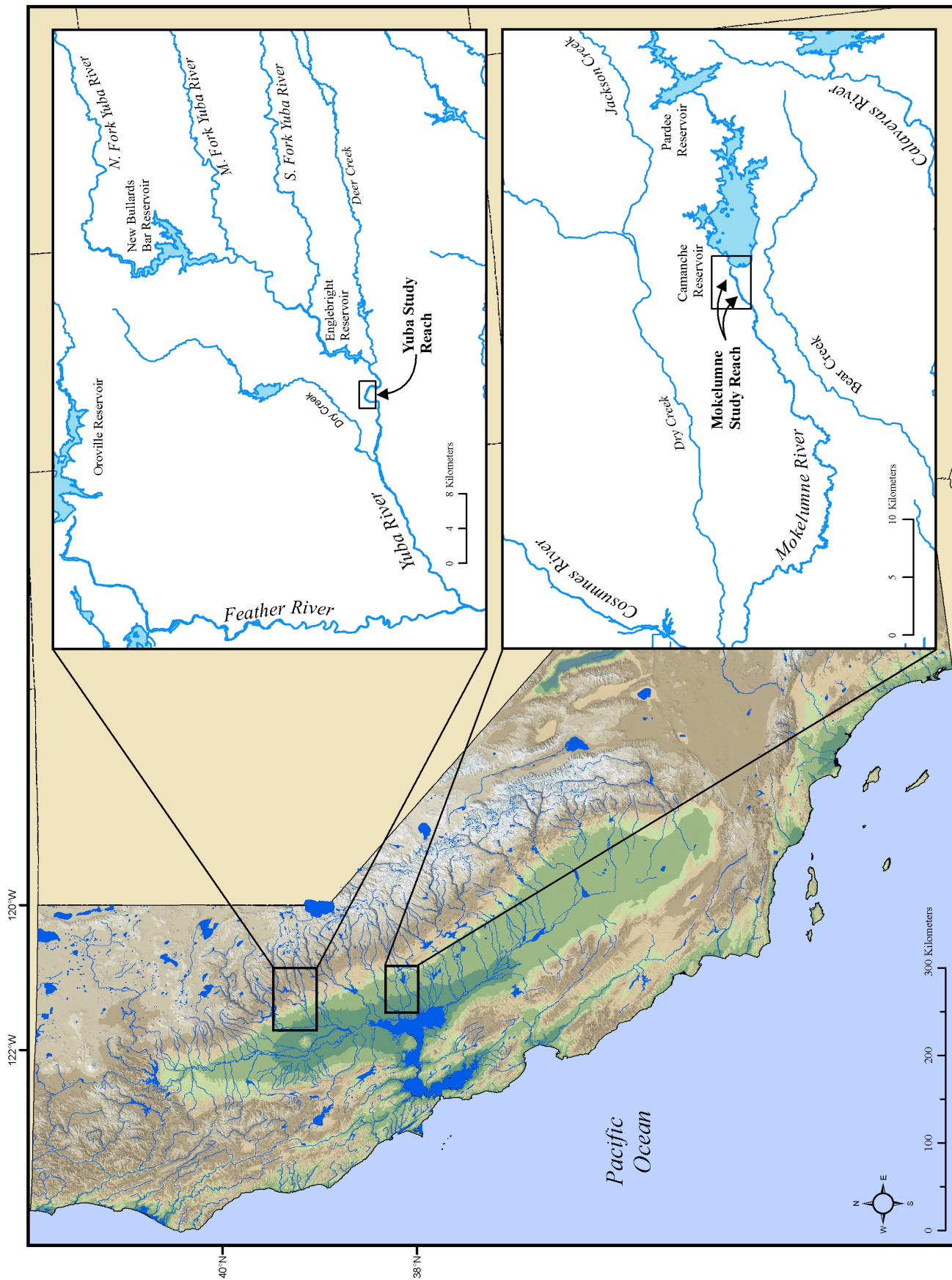
p-level	rQ*		#DFF		%Eff	
	p < .05	p < .01	T	T	p < .025	T
	KS test					
S	p < .001		T	T	T	T
D ₅₀	p < .001		T	T	T	T
c	p < .001		T	T	T	T
f	p > .10		F	F	F	F

Watershed
Scale

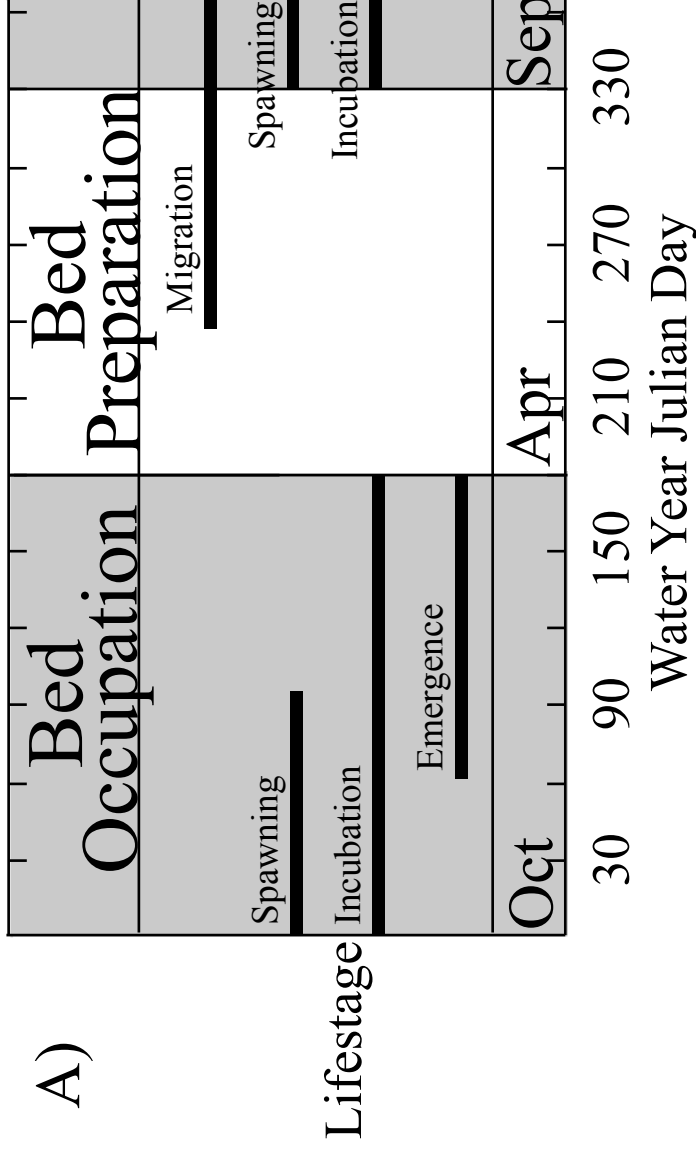


Reach/
Habitat-Unit
Scale

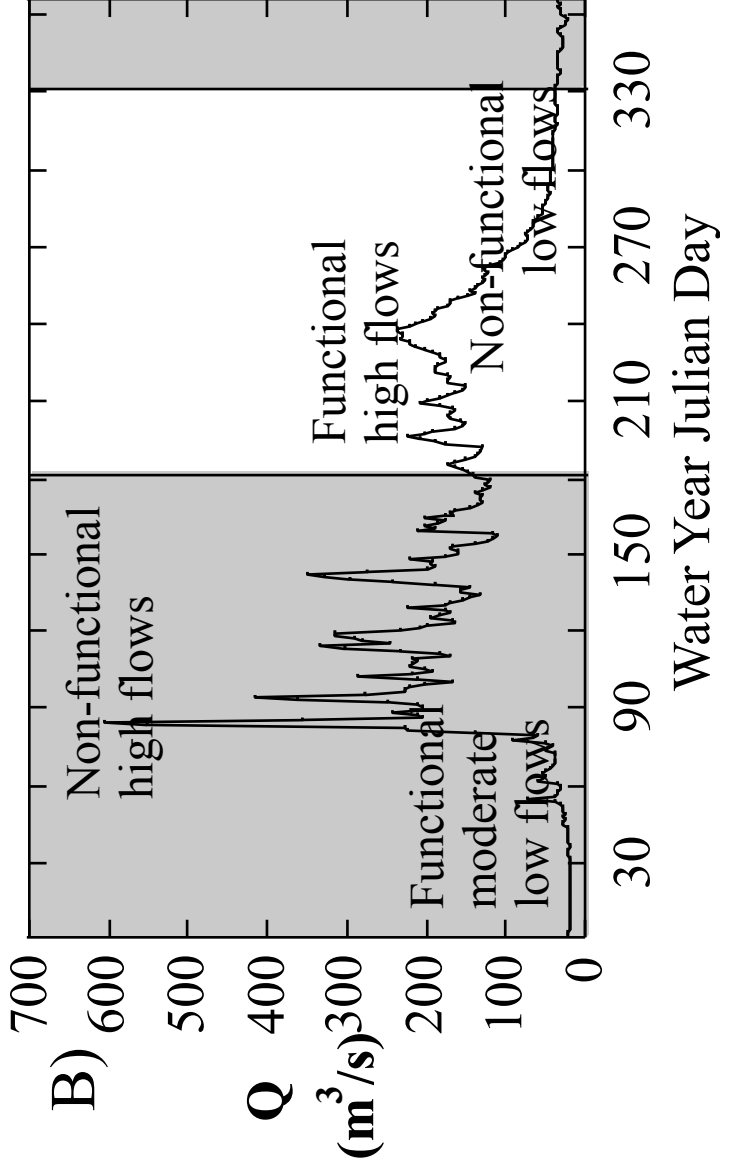


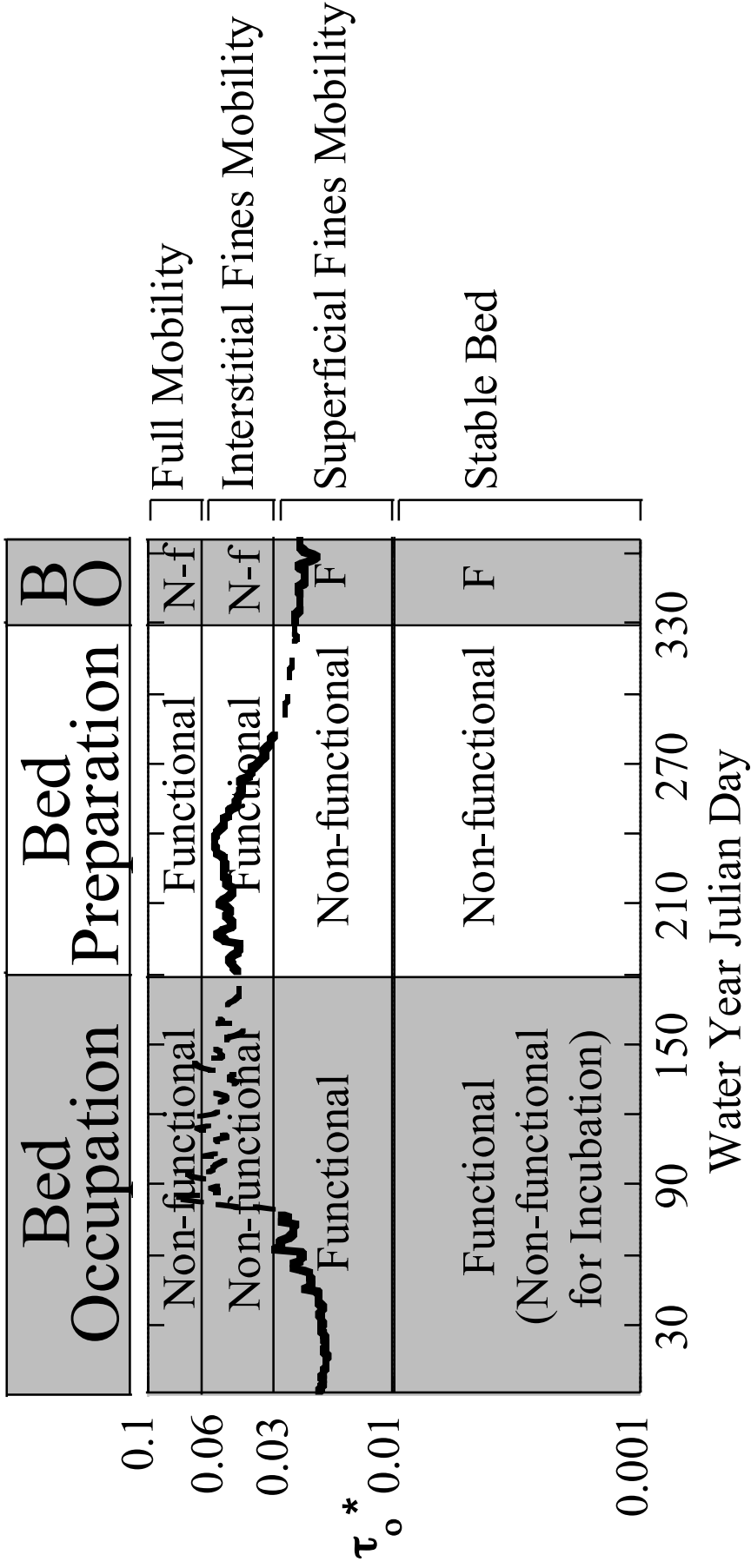


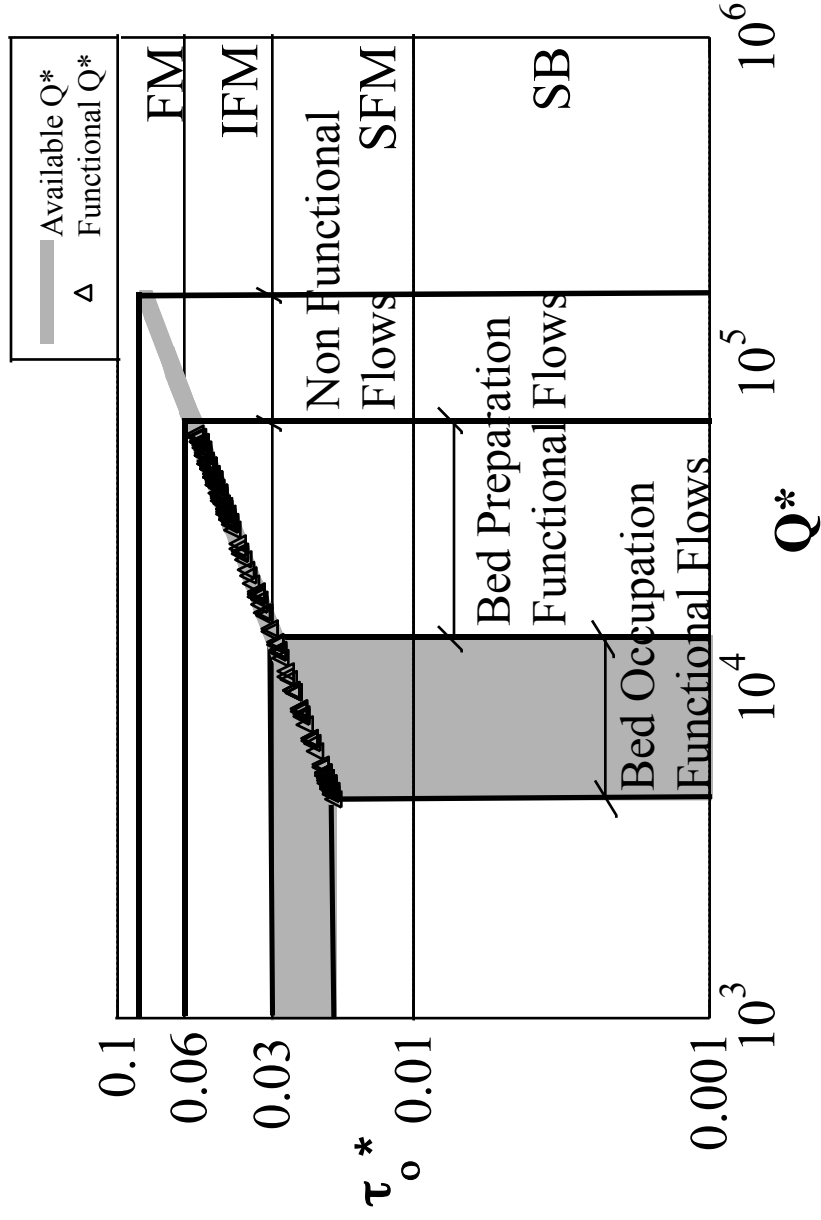
A)

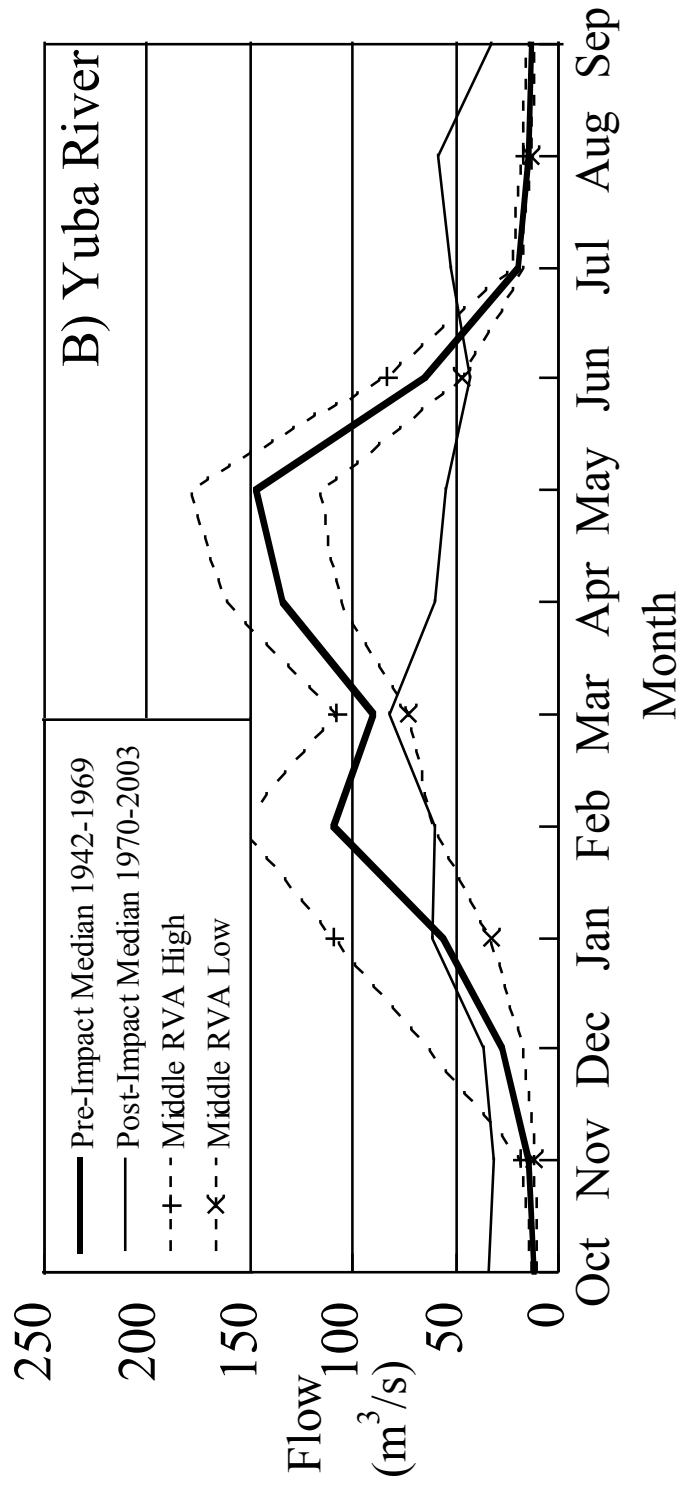
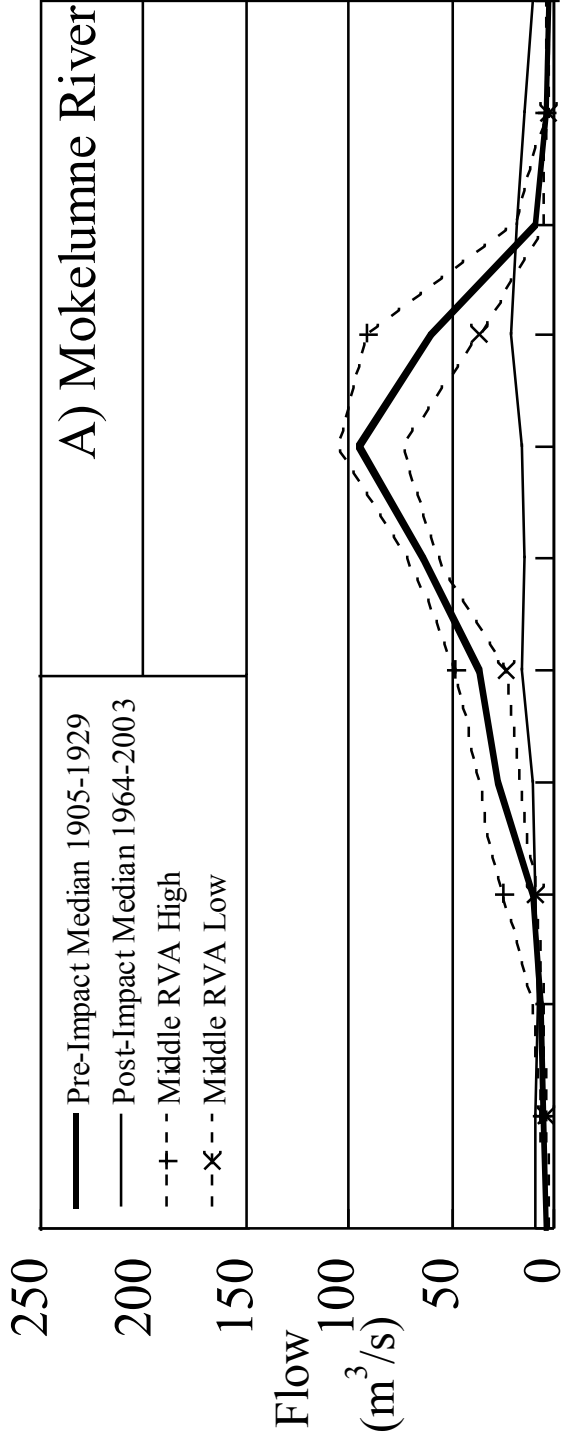


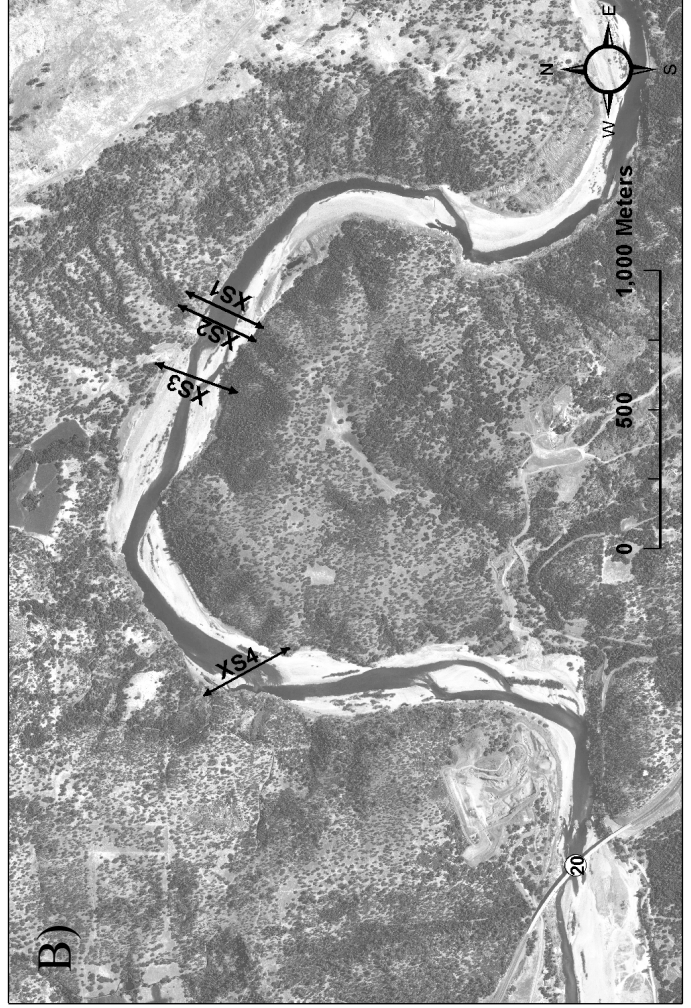
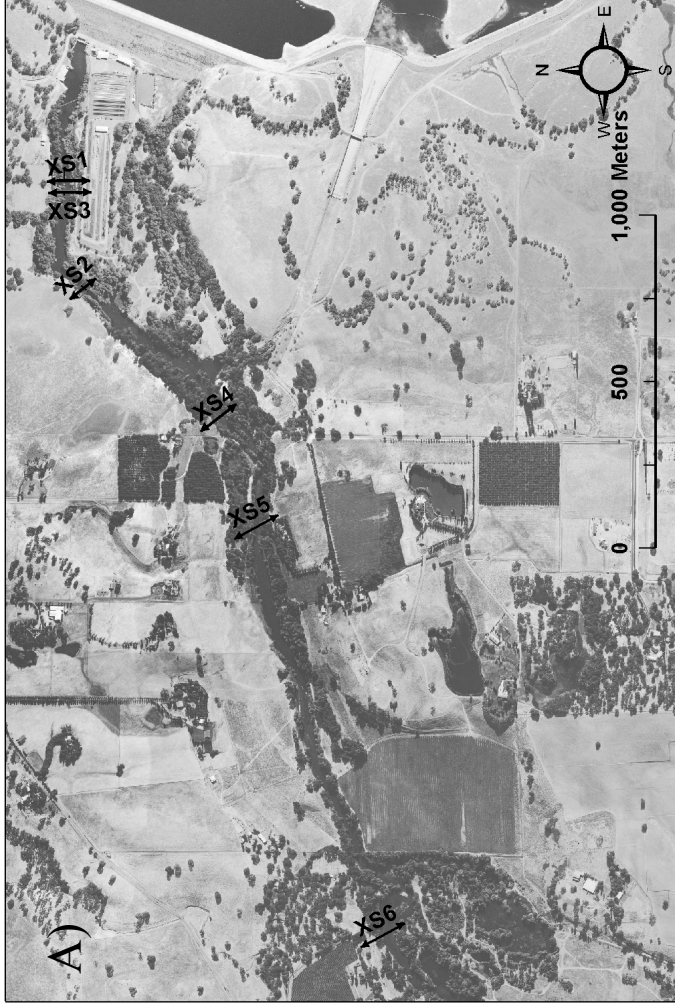
B)

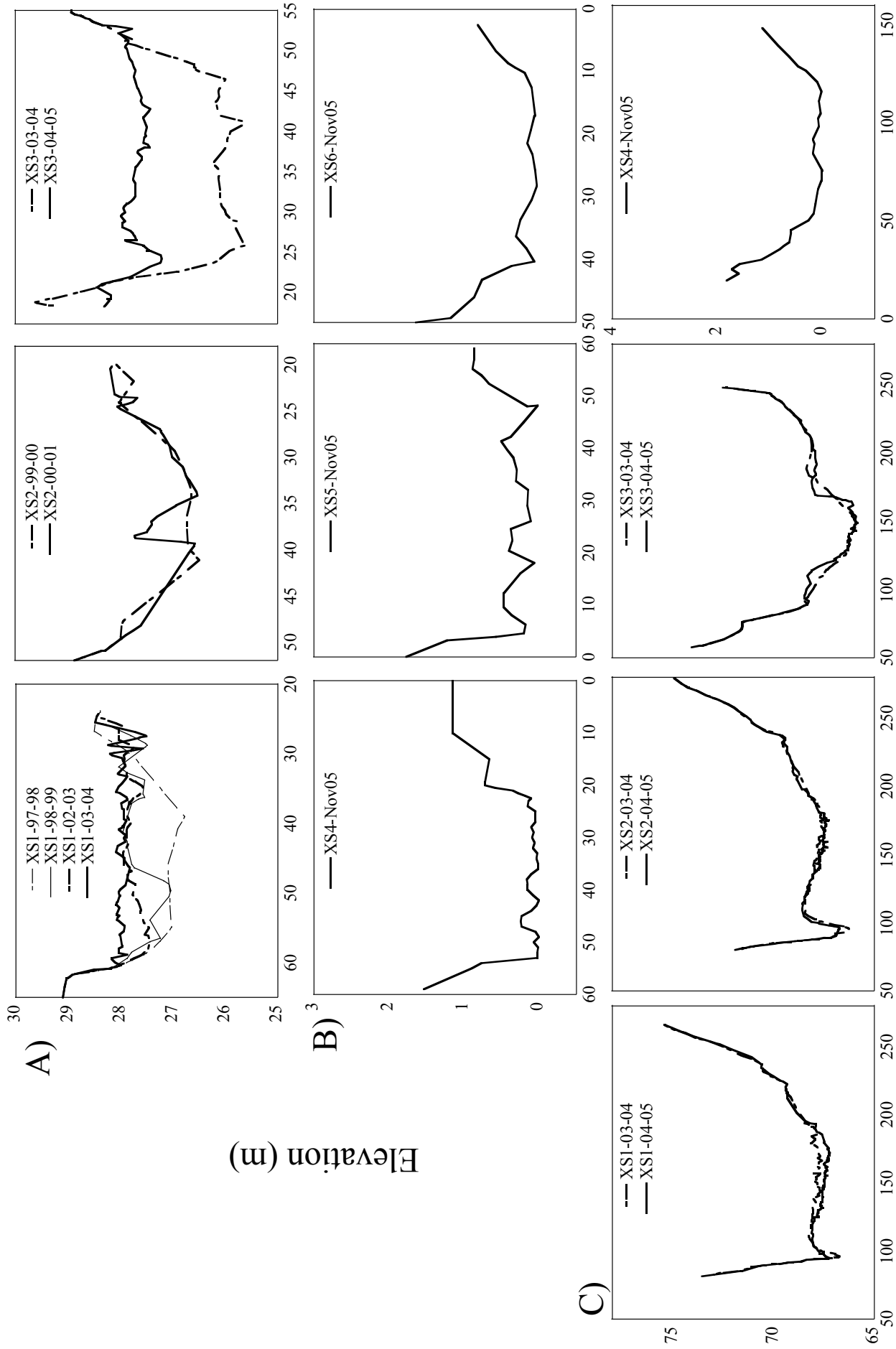












Distance from left bank (m)

